

UNIVERSIDADE FEDERAL DE ALFENAS

UNIFAL- MG

**LAURA FORNERO GOMES FERREIRA**

**INFLUÊNCIA DA ESTRUTURA DA PAISAGEM SOBRE ASPECTOS  
REPRODUTIVOS DE UMA PLANTA FLORESTAL**

ALFENAS – MG  
2013

**LAURA FORNERO GOMES FERREIRA**

**INFLUÊNCIA DA ESTRUTURA DA PAISAGEM SOBRE ASPECTOS  
REPRODUTIVOS DE UMA PLANTA FLORESTAL**

Dissertação apresentada como parte dos requisitos para obtenção do título de Mestre em Ecologia e Tecnologia Ambiental pela Universidade Federal de Alfenas. Área de concentração: Meio Ambiente Sociedade e Diversidade Biológica. Orientador: Prof. Dr. Flavio Nunes Ramos, Co-orientador (a): Prof. Dra. Erica Hasui

ALFENAS – MG  
2013

Ferreira, Laura Fornero Gomes.

Influência da estrutura da paisagem sobre aspectos reprodutivos de uma planta florestal. / Laura Fornero Gomes Ferreira. - 2013. 86 f. -

Orientador: Flavio Nunes Ramos.

Dissertação (Mestrado em Ecologia e Tecnologia Ambiental) - Universidade Federal de Alfenas, Alfenas, MG, 2013.

Bibliografia.

1. Mata Atlântica. 2. Polinização. 3. Floração. I. Ramos, Flavio Nunes. II. Título.

CDD: 577



MINISTÉRIO DA EDUCAÇÃO  
Universidade Federal de Alfenas / UNIFAL-MG  
Programa de Pós-graduação – Ecologia e Tecnologia Ambiental

Rua Gabriel Monteiro da Silva, 714. Alfenas - MG CEP 37130-000  
Fone: (35) 3299-1419 (Coordenação) / (35) 3299-1392 (Secretaria)  
www.unifal-mg.edu.br/ppgecoambiental/



**LAURA FORNERO GOMES FERREIRA**

**“INFLUÊNCIA DA ESTRUTURA DA PAISAGEM SOBRE ASPECTOS REPRODUTIVOS DE UMA PLANTA FLORESTAL”.**

A Banca examinadora, abaixo assinada, aprova a Dissertação apresentada como parte dos requisitos para a obtenção do título de Mestre em Ecologia e Tecnologia Ambiental pela Universidade Federal de Alfenas. Área de Pesquisa: Meio Ambiente, Sociedade e Diversidade Biológica.

Aprovado em: 27/01/2013

Prof. Dr. Flavio Nunes Ramos  
Instituição: UNIFAL-MG

Assinatura:

Prof. Flavio Antonio Maës dos Santos  
Instituição: UNICAMP

Assinatura:

Prof. Dr. Milton Cezar Ribeiro  
Instituição: UNESP

Assinatura:

Dedico esse trabalho ao vô Tino, aos meus pais, Walter e Maria Imaculada, ao meu irmão, Leonardo e ao Alexandre.

## AGRADECIMENTOS

À DEUS, por sempre iluminar meus caminhos.

Aos meus pais Walter e Maria Imaculada e ao meu irmão Leonardo, por serem minha estrutura e me ajudarem a vencer cada desafio.

Ao meu namorado Alexandre, pelo carinho, paciência e pelo apoio em todos os momentos.

Ao meu orientador, Flavio Nunes Ramos, pelo conhecimento passado e paciência. E por ter caído no “caldeirão como o Obelix” ficando assim sempre empolgado e disposto a ajudar.

À professora Erica pela coorientação, motivação e valiosas colaborações e ao professor Rogério Grassetto pela convivência e amizade.

Às amigas Danielle Cavalcante, Cristiane Lopes, Helga Andrade e Grasielle Evangelista que apesar de estarem longe sempre dão um jeitinho de estarem presentes em minha vida, me apoiando.

Às amigas Josiane Pires, Pórtya Cavalcanti, Magda Carneiro, Mariane, Ana Carolina e Driéli pelas companhias e apoio.

Às colegas Gabriela Carnevalli, Marcela Alcântara pela ajuda com o trabalho de campo e dedicação.

Aos colegas Davi Ferri, Marco Túlio Pacheco, Bruno Ribeiro e Mariana Raniero pelo auxílio nas análises do trabalho.

Aos demais amigos e familiares.

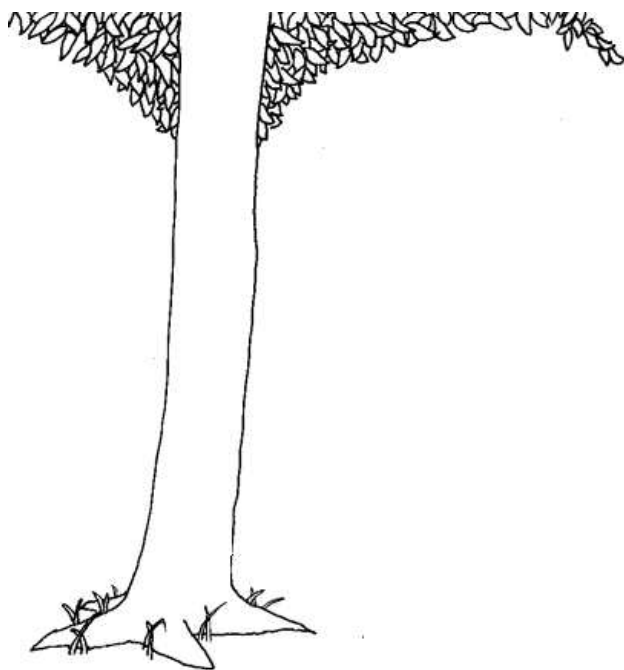
À Universidade Federal de Alfenas e Programa de Pós Graduação em Ecologia e Tecnologia Ambiental pela oportunidade.

A todos os professores e alunos do Laboratório de Ecologia de Fragmentos Florestais (ECOFRAG) pelo aprendizado.

À Coordenação de Aperfeiçoamento de Pessoal de Nível Superior (CAPES) pela bolsa concedida.

À banca avaliadora pelas considerações.

**MUITO OBRIGADA!**



“É do buscar e não do achar que nasce o que eu não conhecia.”

*Clarice Lispector, adaptado.*



## RESUMO

Perda de habitat e mudanças na estrutura da paisagem são efeitos resultantes da fragmentação florestal e podem afetar de maneira negativa processos ecológicos, como a reprodução de plantas e polinização. Nossa hipótese é que parâmetros da estrutura da paisagem (área do fragmento, proximidade, porcentagem de cobertura vegetal da paisagem – floresta nativa, matriz de café, cana e pastagem e formato do fragmento) e características estruturais da planta (circunferência na altura do solo, circunferência na altura do peito, altura da planta, volume da copa) e do seu habitat (altura do dossel, perfil vertical, distância do indivíduo mais próximo e distância do indivíduo coespecífico) afetam a produção de flores e frutos e o número de visitas florais de uma planta florestal. Nós amostramos 160 indivíduos de uma planta florestal, *Psychotria vellosiana*, em oito paisagens (raio de 1000m), cada uma dessas paisagens centrada em um fragmento florestal (20 indivíduos por fragmento). Aplicamos o modelo linear generalizado e critério de informação de Akaike, para selecionar os melhores modelos que explicam as relações entre a reprodução das plantas com: i) os parâmetros da paisagem e ii) medidas de características estruturais das plantas e do seu habitat. Nossos resultados indicam que o número de visitas florais e a produção de flores e frutos são afetados pelo tipo de matriz. Além disso, o número total de visitas florais é influenciado pela porcentagem de floresta concomitantemente à porcentagem de matriz de café. A produção de frutos também é afetada pela complexidade da paisagem, altura do indivíduo e pelo grau de agregação da população de plantas. A partir desses resultados, nós podemos concluir que características da paisagem são importantes para explicar a reprodução das plantas. As informações geradas poderão contribuir para o preenchimento da lacuna no conhecimento a respeito da influência da matriz sobre plantas e processos ecológicos. Além de poderem ser usadas em recomendações destinadas a produtores rurais, tomadores de decisão e políticas públicas, como por exemplo, para informar qual tipo de matriz deve ser usada em paisagens fragmentadas para ajudar a manter ou aumentar a produção de frutos de plantas florestais.

Palavras chave: fragmentação, cobertura florestal, matriz, sucesso reprodutivo das plantas, polinização.

## ABSTRACT

Habitat loss and changes in the landscape structure are effects that results of the forest fragmentation and can negative affect ecological processes, as plant reproduction and pollination. We hypothesized that the landscape structure parameters (fragment area, proximity, percentage of landscape vegetation cover - native forest, coffee, sugarcane and pasture matrix, and fragment shape) and structural characteristics of plant (height stem circumference, circumference at breast height, plant height and volume of crown) and its habitat characteristics (canopy height, vertical profile, distance to the nearest individual and distance to the nearest conspecific individual) affect the flower and fruit production and number of floral visits of forest plant. We sampled 160 individuals of a forest plant, *Psychotria vellosiana*, in eight landscapes (1000m radius), centered each one in one fragment, with 20 individuals per fragment. And we applied a general linear model and Akaike information criterion to select the best models that explains relations among plant reproduction with i) landscape and patch parameters and ii) measurements of structural characteristics of the plants and its habitat. We found that the number of floral visits and flower and fruit production are affected by the matrix type. Furthermore the total number of floral visits is influenced by the percentage of forest concomitantly with percentage of coffee matrix. And fruit production is affected by landscape complexity, size of individual plant and degree of aggregation of the plant population. From these findings, we may conclude that landscapes characteristics are important to explain the plant reproduction. The information generated may contribute to the theoretical gap in knowledge on the influence of matrix on plants and on ecological process. In addition, they can be used in recommendations to farmers, decision makers and public policies, i.e. to inform what type of matrix should be used in fragmented landscape to help to maintain or improve the fruit production of forest plants.

Keywords: fragmentation, forest cover, matrix, plant fitness, pollination.

## SUMÁRIO

<b>1 INTRODUÇÃO GERAL</b> .....	11
1.1 PREÂMBULO.....	11
1.2 PARÂMETROS ESTRUTURAIS DE PAISAGEM E SUA INFLUÊNCIA SOBRE PROCESSOS REPRODUTIVOS DAS PLANTAS.....	11
1.2.1 Área do fragmento.....	12
1.2.2 Formato dos fragmentos.....	13
1.2.3 Distância entre fragmentos.....	15
1.2.4 Matriz.....	16
1.2.5 Características locais.....	18
<b>2 JUSTIFICATIVA</b> .....	19
<b>3 REFERÊNCIAS BIBLIOGRÁFICAS</b> .....	21
<b>SEGUNDA PARTE</b> .....	31
<b>ARTIGO: PROPORTION OF HABITAT, TYPE OF MATRIX AND FRAGMENT SHAPE INFLUENCE THE PLANT REPRODUCTION</b> .....	31
<b>1 INTRODUCTION</b> .....	33
<b>2 METHODS</b> .....	36
2.1. STUDY AREA AND LANDSCAPE SELECTION .....	36
2.2. <i>Psychotria vellosiana</i> AS SPECIES MODEL.....	37
2.3. QUANTIFICATION OF LANDSCAPE STRUCTURE.....	38
2.4. PLANT AND HABITAT CHARACTERISTICS.....	39
2.5. PHENOLOGY.....	40
2.6. FLORAL VISITS.....	41
2.7. FRUIT AND SEED SET.....	42
2.8. DATA ANALYSIS.....	42
2.8.1. Relation of flower and fruit production with traits of plant and their habitats.....	42
2.8.2. Relation of plant reproduction parameters with landscape metrics.....	43
<b>3 RESULTS</b> .....	44
<b>4 DISCUSSION</b> .....	45

<b>5 CONCLUSION</b> .....	49
ACKNOWLEDGEMENTS.....	50
REFERENCES.....	50
<b>4 CONSIDERAÇÕES FINAIS</b> .....	80
<b>5 ANEXO</b> .....	83

# 1 INTRODUÇÃO GERAL

## 1.1 PREÂMBULO

A presente dissertação avalia o efeito da estrutura da paisagem e de características locais (planta e habitat) sobre processos ecológicos (reprodução das plantas e polinização) em uma região altamente fragmentada. Essas relações ainda são pouco exploradas na literatura, principalmente no que diz respeito à resposta das plantas ao tipo de matriz. Como em locais como o estudado, na região de Mata Atlântica, a faixa da quantidade de habitat se encontra em níveis extremamente críticos, torna-se importante entender o que está acontecendo com as espécies que estão sobrevivendo neste ambiente, a fim de gerar conhecimento e informações que ajudem em questões de conservação e restauração.

## 1.2 PARÂMETROS ESTRUTURAIS DE PAISAGEM E SUA INFLUÊNCIA SOBRE PROCESSOS REPRODUTIVOS DAS PLANTAS

A Ecologia de Paisagens é uma ciência recente, surgida na Europa entre 1930 e 1940. Sua abordagem inicial foi impulsionada essencialmente por pesquisadores geógrafos e focava na percepção, uso e ordenamento do espaço de vida do homem (NAVEH, 1982; TURNER, 2001). Após o surgimento de duas teorias pioneiras que relacionavam padrões espaciais e processos ecológicos, a Teoria do Equilíbrio Dinâmico em Ilhas (MACARTHUR; WILSON, 1967) e a Teoria de Metapopulações (LEVINS, 1969), a Ecologia de Paisagem recebeu um enfoque mais biocêntrico (METZGER, 2001; RICKETTS, 2001; THEOBALD, 2006; TURNER, 2005; WIENS; MOSS, 2005;). E dentro deste contexto, pode ser definida sob a ótica das duas abordagens: a geográfica e a ecológica. A geográfica enfatiza o estudo da influência antrópica sobre a paisagem e a gestão do território, e a ecológica destaca a importância do contexto espacial sobre os processos ecológicos e a relevância destas relações no contexto de conservação biológica (METZGER, 2001a; PIVELLO; METZGER, 2007).

Pode se afirmar que, a Ecologia de Paisagem tem como foco principal uma unidade espacializada, heterogênea, composta de um mosaico de unidades interativas (unidades de uso

e ocupação da terra, cobertura vegetal) e integradora de características físicas, biológicas e antrópicas em determinada região - a paisagem (METZGER, 1999; TROPMAIR, 2000). A estrutura da paisagem pode ser definida pela área, forma e disposição espacial (grau de fragmentação e de isolamento, por exemplo) (FORMAN, 1995; METZGER, 2001; OPDAM et al., 1993) das unidades interativas.

Processos antrópicos como a fragmentação de habitat rompem a continuidade das unidades interativas da paisagem (ecossistemas, unidades de vegetação) (LORD; NORTON, 1990) e modificam sua estrutura (METZGER, 1999). Essa mudança na estrutura da paisagem pode resultar em alterações no microclima, na sobrevivência e deslocamento das espécies, na composição e diversidade das comunidades e conseqüentemente, em processos ecológicos, como a reprodução vegetal (BREITBACH et al., 2012; GOULSON et al., 2008; METZGER, 1999). Assim, acredita-se que alterações em parâmetros de paisagem, como diminuição na área de habitat (área do fragmento, cobertura de vegetação nativa), modificações na forma do remanescente, no grau de isolamento entre eles, além do tipo de matriz circundante, podem influenciar de forma direta ou indireta a reprodução das plantas (polinização, produção de frutos e sementes) (BREITBACH et al., 2012; FISCHER, LINDENMAYER, 2007; JULES; SHAHANI, 2003).

### 1.2.1 Área do fragmento

A fragmentação florestal resulta na subdivisão de áreas contínuas de floresta nativa em manchas ou fragmentos florestais menores e isolados (WILCOVE et al., 1986). Concomitantemente, à redução de tamanho de áreas nativas pode ocorrer a perda de habitat e assim um depauperamento dos processos ecológicos, como a reprodução das plantas. Como a área do fragmento fica menor do que as áreas mínimas, necessárias para a sobrevivência das populações de plantas e polinizadores (FORMAN et al., 1976; SAUNDERS et al., 1991), pode ocorrer uma diminuição da abundância, riqueza e biodiversidade desses organismos bem como mudanças em suas distribuições (capacidade de dispersão) (GIBBS, 1998; GUTHERY et al., 2001; STEFFAN-DEWENTER et al., 2002). Alterações no comportamento dos polinizadores, que podem prejudicar as taxas de sucesso de forrageamento (MAHAN; YAHNER, 1999) também são esperadas com a redução do tamanho do fragmento, além de redução do número de plantas compatíveis (devido a menor densidade de plantas e flores) nos

fragmentos (AIZEN; FEINSINGER, 1994a; GRIFFIN; ECKERT, 2003; WAGENIUS et al., 2007). Todas essas consequências negativas podem induzir a limitação de pólen e consequentemente afetar a polinização e o sucesso reprodutivo das plantas (AIZEN; FEINSINGER, 1994a, 1994b; FAHRIG, 2003; KURKI et al., 2000; VAMOSI et al., 2006; VIANA et al., 2012).

Algumas espécies são mais sensíveis à redução da área do fragmento e à mudanças microclimáticas, enquanto outras são mais tolerantes a espaços abertos, como pastagens e áreas urbanas (KRAUSS et al., 2003; WETHERED; LAWES, 2003). As abelhas, por exemplo, respondem de maneira diferente ao efeito da fragmentação de acordo com a espécie. Nos projetos “Insetos e Aves” e “Restauna”, que estudaram os efeitos da fragmentação sobre grupos de visitantes florais, foi verificado uma menor riqueza de espécies de abelhas sem ferrão nos fragmentos menores, já que algumas espécies de abelhas desse grupo são encontradas somente em fragmentos maiores (ex. *Melipona quadrifasciata* e *Trigona fulviventris*). A presença de outras abelhas sem ferrão (*Melipona quadrifasciata*) estava condicionada à estrutura vegetacional (riqueza de espécies e altura de árvores); ao ambiente de interior dos fragmentos (*Cephalotrigona captata*); ou ainda a áreas não urbanizadas (*Geotrigona subterrânea*). Ou seja, abelhas da espécie *G. subterrânea*, que fazem ninhos no solo, de preferência não impermeabilizado, são encontradas dentro e fora de fragmentos, desde que estes estejam em áreas não urbanizadas (RAMBALDI; OLIVEIRA, 2005).

Já as abelhas da subtribo Euglossina, nesse mesmo estudo, tenderam a apresentar menor riqueza de espécies quando a área central (subtraída a área de borda) dos fragmentos da Região Metropolitana de Belo Horizonte foram reduzidas. Quando esses fragmentos menores estavam próximos (menos de 500m) de outros maiores, essa tendência foi amenizada e uma riqueza maior de espécies foi encontrada nos remanescentes pequenos (RAMBALDI; OLIVEIRA, 2005). Dentro desse contexto, fica evidente que a área do fragmento é um dos parâmetros mais importantes para explicar variações na riqueza, diversidade das espécies e nas interações ecológicas (por exemplo, inseto planta). Logo, a área pode ser considerada uma métrica relevante para estudos dos efeitos da estrutura da paisagem sobre a reprodução das plantas.

### 1.2.2 Formato dos fragmentos

A forma do fragmento interfere na proporção da sua área de borda (METZGER, 2003), sendo que quanto mais alongado e irregular ele for maior será o efeito de borda (MCGARIGAL; MARKS, 1995; MURCIA, 1995). O aumento da borda, devido à fragmentação florestal, pode fazer com que a área central do fragmento se torne menos protegida de fatores externos e fique mais próxima à matriz circundante (ambientes estruturalmente diferentes das florestas), o que pode causar modificações nas condições abióticas do fragmento (MCGARIGAL; MARKS, 1995; METZGER, 1999; MURCIA, 1995). Essas modificações abióticas podem estimular mudanças na estrutura florestal da borda, na composição de espécies do fragmento e nas interações ecológicas (MURCIA, 1995).

As alterações na estrutura florestal da borda podem ocorrer porque mudanças microclimáticas são capazes de interferir no crescimento, na mortalidade, na abundância e na distribuição das plantas deste novo ambiente (MURCIA, 1995). Já as modificações na composição de espécies são esperadas, porque o aumento da área de borda pode alterar a permeabilidade do fragmento a espécies exóticas (aumento da entrada), o que pode representar uma ameaça às espécies originalmente presentes nos remanescentes (HILL; CURRAN, 2001; SAUNDERS et al., 1991; TURNER; CORLETT, 1996). Além disso, com aumento do efeito de borda são esperadas modificações em interações ecológicas (MURCIA, 1995), como na polinização (AIZEN; FEINSINGER, 1994a; SAUNDERS et al., 1991) pois, algumas espécies de polinizadores, principalmente insetos, são influenciados por variações na temperatura e umidade (HERRERA, 1995), e suas atividades dependem de um microclima apropriado dentro da flor e do ambiente (CORBET, 1990). Essas mudanças, redução de espécies (plantas e polinizadores) e modificações no comportamento dos polinizadores, podem resultar em diminuição das taxas de visitação e da quantidade e qualidade do pólen que chega ao estigma das flores e conseqüentemente num prejuízo no sucesso reprodutivo das plantas (AIZEN; FEINSINGER, 1994a; HERRERA, 2000; WASER; PRICE, 1991).

A partir dos dados acima, pode-se presumir que o formato do fragmento é um parâmetro importante para estudar a influência do contexto da paisagem sobre processos ecológicos, como a reprodução das plantas. E entre os índices que quantificam esse parâmetro, pode-se citar o índice de forma média ponderado pela área (AWMSI) que calcula a complexidade dos remanescentes da paisagem realizando uma média ponderada pela área. O valor do índice é igual a 1 quando todos os remanescentes são circulares e aumenta à medida que as formas dos fragmentos ficam mais complexas e as áreas maiores (MCGARIGAL; MARKS, 1995).



### 1.2.3 Distância entre fragmentos

A distância entre os remanescentes florestais (isolamento) é um dos fatores responsáveis pela: conectividade (capacidade da paisagem de facilitar fluxos biológicos de organismos) (URBAN; SHUGART, 1986) entre os fragmentos e habitats não fragmentados e influencia as taxas de dispersão, imigração e recolonização dos remanescentes (HAILA et al., 1993). Imigrantes contribuem demográfica e geneticamente para populações dos fragmentos aumentando o tamanho e a aptidão de suas populações (BROWN; KODRIC-BROWN, 1977). Assim, em geral, é esperado que espécies de plantas e outros organismos em fragmentos menos isolados e/ou em paisagens mais conectadas sofram menos problemas de trocas gênicas e declínio populacional (LAURENCE; VASCONCELOS, 2009; RAMBALDI; OLIVEIRA, 2005).

A dificuldade de movimentação de organismos entre fragmentos é proporcional ao grau de isolamento, ao nível de resistência da matriz circundante e do organismo em questão. Por exemplo, algumas aves podem movimentar-se entre fragmentos mesmo em situações de conectividade reduzida; grandes vertebrados podem encontrar maior resistência e pequenos mamíferos, em geral, demonstram padrões intermediários (RAMBALDI; OLIVEIRA, 2005).

Já o isolamento espacial de espécies de árvores pode modificar a atividade de polinizadores reduzindo a densidade de recursos alimentares e aumentando a distância entre esses recursos (SIH; BALTUS, 1987). A redução de recursos florais resulta em viagens com distâncias mais longas, entre áreas de descanso e alimentação, e frequentemente resulta no deslocamento de polinizadores por áreas menos favoráveis, por exemplo, áreas dominadas por agricultura (KEARNS et al., 1998). Quando a distância entre plantas é maior que a área de vida de polinizadores, a densidade deles diminuirá nas áreas que sofreram distúrbios e isto resultará em menos visitas (KEARNS et al., 1998). Polinizadores especialistas, que são menos flexíveis na exploração de recursos alimentares são os mais suscetíveis à extinção local (COSSON et al., 1999; ESTRADA et al., 1993; STONER et al., 2002). Logo, o isolamento entre remanescentes pode influenciar a diversidade de polinizadores e seu comportamento de forrageamento e conseqüentemente o sucesso reprodutivo e sistemas de cruzamento das plantas que eles polinizam (AIZEN; FEINSINGER, 1994; FUCHS et al., 2003; QUESADA et al., 2003). Portanto, a distância entre fragmentos pode ser considerada um parâmetro importante para ajudar a explicar mudanças na reprodução das plantas devido a mudanças no contexto da paisagem.

#### 1.2.4 Matriz

Matrizes são áreas que antes eram cobertas por vegetação natural e agora são mosaicos de habitats modificados pelo homem que circundam os fragmentos florestais, por exemplo: pastagens, plantações, florestas em regeneração, estradas e áreas urbanas (WIENS et al., 1985). Antes, a matriz era considerada sem relevância ecológica, pois era vista como homogênea e desabitada (JULES; SHAHANI, 2003; RICKETTS, 2001). Contudo, após vários estudos, verificou-se que ela é um elemento importante da paisagem, pois pode influenciar relações tróficas, habitat, movimentação de organismos, reprodução e forrageamento.

Diferentes tipos de matriz podem influenciar na qualidade do fragmento florestal devido à (i) condições abióticas: por exemplo, as características topográficas, edáficas, microclimáticas de cada tipo de matriz, podem determinar a composição e distribuição de plantas nos fragmentos (WIENS et al., 1985); (ii) ambiente de transição: plantas da matriz podem interagir com plantas da borda do fragmento ao competir por luz, água, nutrientes e espaço ou funcionar como atração adicional a animais como alguns dispersores e polinizadores. A matriz também pode influenciar o fluxo de água, vento e fogo no interior dos fragmentos (GASCON et al., 2000; JULES; SHAHANI, 2003; PICKETT; CADENASSO, 1995; WEATHERS et al., 2001; WIENS et al., 1985). O fogo iniciado na matriz pode penetrar os fragmentos e diminuir a cobertura vegetal, aumentar as áreas de solo exposto e provocar a morte das plantas das bordas adjacentes, alterando a estrutura e a dinâmica das populações de plantas e outros organismos (COCHRANE; LAURANCE, 2002; GASCON et al., 2000; SANAIOTTI; MAGNUSSON, 1995; SCARIOT et al., 2005).

Além disso, diferentes usos da matriz podem estar relacionados ao aumento ou diminuição das áreas dos fragmentos (HAYNES et al., 2007), à conectividade entre eles (BENDER; FAHRIG, 2005), à utilização pelos organismos de corredores e stepping-stones (áreas de vegetação que podem permitir o fluxo de indivíduos) (BAUM et al., 2004) e à natureza da borda (HODGSON et al., 2007). Quanto maior a semelhança da matriz ao habitat do remanescente, maior a conectividade funcional (coesão entre populações de diferentes remanescentes) (ANDERSON et al., 2007; MEDELLÍN et al., 2000; RICKETTS, 2001; TAYLOR et al., 1993). Logo, cada tipo de matriz pode influenciar de maneira diferente movimentos de polinizadores, dispersores de sementes e herbívoros (JULES; SHAHANI, 2003).

O tipo de vegetação da matriz é responsável pelo tamanho do poro do filtro no movimento dos indivíduos (MURPHY; LOVETT-DOUST, 2004). Enquanto florestas secundárias permitem maior movimentação de organismos entre os remanescentes, por possuir a estrutura da vegetação mais semelhante à dos fragmentos, as matrizes de pasto podem impedir grande parte desse movimento (GASCON et al., 1999). Assim, acredita-se que algumas matrizes são mais favoráveis a fauna da floresta do que outras. No Amazonas, observou-se que florestas secundárias dominadas por *Cecropia* sp (STOUFFER; BIERREGAARD, 1995) e *Cecropia sciadophylla* Mart. (STOUFFER et al., 2006) foram mais utilizadas por espécies de pássaros (STOUFFER et al., 2006; STOUFFER; BIERREGAARD, 1995), sapos (TOCHER, 1998) e formigas, do que as florestas secundárias dominadas por *Vismia* (exemplo: *V. guianenses* (Aubl.) Choisy, *V. japurensis* Reich. e *V. cayennensis* (Jacq.) Pers.) que são mais baixas e estruturalmente menos diversa (MESQUITA et al., 2001). Outro trabalho mostrou que a abundância total de pequenos mamíferos foi maior em matriz antrópica com cobertura florestal (plantações de eucalipto) do que nas sem cobertura florestal (áreas de agricultura e área rural com construções) (UMETSU, 2005).

Matrizes de diferentes composições geralmente possuem microclimas distintos (maior temperatura e luminosidade, menor umidade) em relação aos fragmentos. Desta forma, acredita-se que elas possam afetar as fenofases das plantas, como por exemplo, a floração, que depende diretamente da quantidade de luz e a frutificação, que está relacionada à disponibilidade de polinizadores que transitam pelas matrizes para ter acesso aos remanescentes (JULES; SHAHANI, 2003). Plantas da matriz, também podem proporcionar uma atração adicional, que pode aumentar ou diminuir (devido à competição) a quantidade de polinizadores e taxas de polinização das plantas dos fragmentos adjacentes (JULES; SHAHANI, 2003; QUINTERO, 2010). Matrizes diferentes, portanto, poderão interferir no sucesso reprodutivo de plantas em fragmentos através de dois caminhos. Primeiro, ao modificar variáveis abióticas, e segundo, ao afetar a sobrevivência, a abundância e movimentação de visitantes florais e a polinização (AIZEN; FEINSINGER, 1994; HERRERA, 2000; RAMBALDI; OLIVEIRA, 2005).

Muitos estudos apontaram os efeitos negativos e deletérios das matrizes antrópicas sobre populações de plantas e animais (MONTERO-CASTAÑO; VILÀ, 2012; KEARNS et al., 1998), enquanto outros indicam que elas podem apresentar consequências positivas (JULES; SHAHANI, 2003). Por exemplo, em um estudo sobre a reprodução de uma planta, verificou-se que efeitos negativos do aumento do número de predadores de sementes devido à matriz foram compensados pelo acréscimo da visita dos polinizadores (STEFFAN-

DEWENTER et al., 2002). Outros fatores importantes a serem levados em consideração dentro do contexto de matrizes são as mudanças temporais (JULES; SHAHANI, 2003). A maior parte dos remanescentes florestais não é circundada por um tipo “fixo” de matriz, mas por habitats em constantes conversões, tanto por influências antrópicas quanto naturais. Pode ocorrer uma conversão de uma área em abandono para agricultura ou pastagem e posteriormente reconversão para abandono ou regeneração de florestas (CARROLL et al., 1990; RUDEL; HOROWITZ, 1993). As mudanças temporais na matriz são importantes porque a qualidade da matriz, que tem potencial de influenciar a dinâmica dos fragmentos, pode ser alterada à medida que a matriz muda. Assim, a permeabilidade da matriz a polinizadores, dispersores, também pode variar no tempo e ocasionar mudanças na reprodução de algumas plantas (JULES; SHAHANI, 2003).

Como a matriz pode influenciar a diversidade, sobrevivência e a reprodução de muitas espécies, é relevante incluí-la em estudos de fragmentação e paisagem, e a partir desses, contribuir para a criação de métodos que diminuam o impacto no ecossistema, causado pelo uso do solo (RICKETTS, 2001; LAURANCE, 2008).

### 1.2.5 Características locais

Além dos parâmetros de paisagem, citados anteriormente, algumas características locais relacionadas à estrutura da planta e ao habitat também podem influenciar a reprodução de plantas florestais (PIZO; ALMEIDA-NETO, 2009). O aumento de medidas estruturais como CAS (circunferência do caule na altura do solo), CAP (circunferência do caule na altura do peito), altura da planta e volume da copa podem influenciar de maneira positiva na reprodução da planta (SOARES, 2011). Assim, plantas pertencentes à mesma espécie, mas com tamanhos distintos, podem apresentar diferenças na produção de flores e frutos, devido à alocação de recursos. Enquanto plantas menores investem seus recursos na produção de raiz e folhas, as plantas maiores alocam os recursos disponíveis na reprodução. Logo, plantas maiores tendem a produzir mais flores, atrair mais polinizadores e conseqüentemente produzir mais frutos (CHARLESWORTH; LEON, 1976; PIÑERO et al., 1982; OSADA et al., 2002).

Características do habitat, relacionadas indiretamente à quantidade de luz que chega à planta, também devem influenciar a reprodução de plantas florestais (SOARES, 2011). Quanto maior a altura do dossel acima da planta, quanto menor o perfil vertical da vegetação

(número de vezes que uma vara de 6m, toca a vegetação acima do indivíduo focal) e menor a distância do indivíduo de planta mais próximo ao estudado, maior a incidência de luz sobre a planta focal. E, por conseguinte, maior a produção de flores e frutos, pois a luz e temperatura são indutores destas fenofases (WRIGHT; VAN SHAIK, 1994).

O grau de agregação da população de plantas também deve afetar a reprodução (BLENDINGER et al., 2008; CARLO; MORALES, 2008). O aumento da distância entre indivíduos coespecíficos pode ter um efeito positivo na produção de flor, devido a maior luminosidade e um efeito positivo ou negativo na produção de frutos (WRIGHT; VAN SHAIK, 1994). Positivo, porque plantas pertencentes à mesma espécie, mas distantes uma da outra, irão competir menos por polinizadores durante a fase de floração, o que pode resultar em maior taxa de visitas florais e produção de frutos (BLENDINGER et al., 2008; RAMOS; SANTOS, 2005). Contudo, esse aumento da distância também pode diminuir a atração de polinizadores, o que pode resultar em menor produção de frutos (BLENDINGER et al., 2008; DUFFY; STOUT, 2011).

Dentro desse contexto, pode-se observar que o estudo da influência de características em escala local, junto a parâmetros em escala de paisagem pode permitir uma avaliação ampla das respostas reprodutiva de plantas a mudanças no habitat causadas pelo homem como a fragmentação (MONTERO-CASTANO; VILA, 2012; PIZO; ALMEIDA-NETO, 2009).

## **2 JUSTIFICATIVA**

As relações abordadas nesta dissertação, efeito da estrutura da paisagem e de características locais (planta e habitat) sobre processos ecológicos (reprodução das plantas e polinização), ainda são pouco exploradas na literatura. E ainda há uma lacuna conceitual, principalmente a respeito da resposta das plantas à matriz da paisagem (FERREIRA, et al., 2013; PREVEDELLO; VIEIRA, 2010). Como a região de estudo está situada no bioma Mata Atlântica, que se encontra extremamente fragmentado (MYERS et al., 2000) e cuja a maioria dos fragmentos (80%) apresentam áreas menores que 50ha (RIBEIRO et al., 2009), torna-se importante estudar os processos ecológicos que ainda ocorrem nessas regiões, a fim de gerar conhecimento e informações que possam ajudar em questões relacionadas a conservação e restauração de áreas florestais fragmentadas, tão comuns na paisagem atual.

Além disso, a utilização da Ecologia de Paisagens como uma ferramenta para investigações das alterações ambientais reforça uma mudança da maneira de conduzir os estudos, ao substituir uma visão prévia de que o meio ambiente é homogêneo e equilibrado por uma visão mais dinâmica e de inter-relações entre padrões espaciais e processos ecológicos (METZGER, 2001).

### 3 REFERÊNCIAS BIBLIOGRÁFICAS

AIZEN, M. A.; FEINSINGER, P. Forest fragmentation, pollination, and plant reproduction in a Chaco dry forest, Argentina. **Ecology**, v.75, p. 33–351, 1994a.

AIZEN, M.A. ; FEINSINGER, P. 1994b. Habitat fragmentation, native insect pollinator, and feral honey bees in Argentina “Chaco Serrano”. **Ecological Applications** 4: 378-392.

ANDERSON, J.; ROWCLIFFLE, J. M.; COWLISHAW, G. Does the matrix matter? A forest primate in a complex agricultural landscape. **Biological Conservation**, v. 135, n. 2, p. 212-222, 2007.

BAUM, K.; HAYNES K. J.; DILLEMUTH, F. P.; CRONIN, J. The matrix enhances the effectiveness of corridors and stepping stones. **Ecology**, v. 85, n. 10, p. 2671–2676, 2004.

BENDER, D. J.; FAHRIG, L. Matrix structure obscures the relationship between interpatch movement and patch size and isolation. **Ecology**, v. 86, n. 4, p. 1023-1033, 2005.

BLENDINGER, P.G.; LOISELLE, B.A.; BLAKE, J. G. Crop size, plant aggregation, and microhabitat type affect fruit removal by birds from individual melastome plants in the Upper Amazon. **Oecologia**, v.158, n. 273–283, 2008.

BREITBACH, N.; TILLMANN, S.; SCHLEUNING, M.; GRÜNEWALD C.; LAUBE, I.; STEFFAN-DEWENTER, I.; BÖHNING-GAESE, K. Influence of habitat complexity and landscape configuration on pollination and seed-dispersal interactions of wild cherry trees. **Oecologia**, v. 168, p. 425–437, 2012.

BROWN, J. H.; KODRIC-BROWN, A. Turnover rates in insular biogeography: effect of immigration on extinction. **Ecology**, v. 58, p. 445-449, 1977.

BURGESS, V. J.; KELLY, D.; ROBERTSON, A. W.; LADLEY, J. J. Positive effects of forest edges on plant reproduction: literature review and a case study of bee visitation to flowers of *Peraxilla tetrapetala* (Loranthaceae). **New Zealand Journal of Ecology**, v.30, n.2, p.1-13, 2006.

CARLO, T.A., MORALES, J.M., 2008. Inequalities in fruit-removal and seed dispersal: consequences of bird behaviour, neighbourhood density and landscape aggregation. **Journal of Ecology**, v. 96, p. 609–618, 2008.

CARROLL, C.; VANDERMEER, J. H.; ROSSET, P. **Agroecology**. New York, NY, US: McGraw-Hill, 1990.

CHARLESWORTH, B., LEON, J.A., 1976. The relation of reproductive effort to age. *American Naturalist*, v.110,p. 449–459, 1976.

COCHRANE, M. A.; LAURANCE, W. F. Fire as a large-scale edge effect in Amazonian forests. **Journal of Tropical Ecology**, v.18, p. 311–325, 2002.

COSSON, J. F.; RINGUET, S.; CLAESSENS, O.; DE MASSARY, A.; DALECKY, J. C.; VILLERS, J. F.; GRANJON, L.; PONS, J. M. Ecological changes in recent land-bridge islands in French Guiana, with emphasis on vertebrate communities. **Biological Conservation**, v. 91, p. 213–222, 1999.

ESTRADA, A.; COATES-ESTRADA, R. Bat species richness and abundance in tropical rain forest fragments and in agricultural habitats at Los Tuxtlas México. **Ecography**, v.16, p. 309–318, 1993.

FAHRIG, L. Effects of habitat fragmentation on biodiversity. *Annual Review of Ecology. Evolution and Systematics*, v. 34, p. 487-515, 2003.

FERREIRA, P. A.; BOSCOLO, B.; BLANDINA, F. V. What do we know about the effects of landscape changes on plant–pollinator interaction networks? **Ecological Indicators**, 2013. in press. Disponível em: <<http://dx.doi.org/10.1016/j.ecolind.2012.07.025>> Acesso em: 9 jan. 2012.

FISHER, J.; LINDENMAYER, D.B. Landscape modification and habitat fragmentation: a synthesis. **Global Ecology and Biogeography**, v. 16, p. 265–280, 2007.

FORMAN, R. T. T. **Land Mosaics: The Ecology of Landscapes and Regions**. Cambridge: Cambridge University Press, 1995.

FORMAN, R. T. T.; GALLI, A. E.; LECK, C. F. Forest size and avian diversity in New Jersey woodlots with some land use implications. **Oecologia**, v. 26, p. 1–8, 1976.

FUCHS, E. J.; LOBO, J. A.; QUESADA, M. Effects of forest fragmentation and flowering phenology on the reproductive success and mating patterns on the tropical dry forest tree, *Pachira quinata* (Bombacaceae). **Conservation Biology**, v. 17, p. 149–157, 2003.



GASCON, C.; LOVEJOY, T. E.; BIERREGAARD, R. O.; MALCOLM, J. R.; STOUFFER, P. C.; VASCONCELOS, H. L.; LAURANCE, W. F.; ZIMMERMAN, B.; TOUCHER, M.; BORGES, S. Matrix habitat and species richness in tropical forest remnants. **Biological Conservation**, v. 91, p. 223-229, 1999.

GASCON, C.; WILLIAMSON, G. B.; DA FONSECA, G. A. B. Receding forest edges and vanishing reserves. **Science**, v. 288, p. 1356–1358, 2000.

GIBBS JP. Distribution of woodland amphibians along a forest fragmentation gradient. **Landscape Ecology**, v.13, p. 263–68, 1998.

GILPIN, M. E.; HANSKI, I. A. **Metapopulation dynamics: empirical and theoretical investigations**. London: Academic Press, 1991.

GOULSON, D.; LYE, G. C.; DARVILL, B. Decline and Conservation of Bumble Bees. **Annual Review of Entomology**, v. 53, p. 191–208, 2008.

GRIFFIN, C. A. M.; ECKERT, C. G. Experimental analysis of biparental inbreeding in a self-fertilizing plant. **Evolution**, v. 57, p. 1513–1519, 2003.

GUTHERY, F. S.; GREEN M. C.; MASTERS R.E.; DEMASO, S. J.; WILSON, H.M.; STEUBING, F.B. 2001. **Landcover and bobwhite abundance on Oklahoma farms and ranches**. *Journal of Wildlife Management*, v. 65, p. 838–49.

HAILA, Y.; HANSKI, I.K.; RAIVIO, S. Turnover of Breeding Birds in Small Forest Fragments: The "Sampling" Colonization Hypothesis Corroborated. **Ecology**, v. 74, p. 714-725, 1993.

HAYNES, K. J.; DIEKÖTTER, T.; CRIST, T. O. Resource complementation and the response of an insect herbivore to habitat area and fragmentation. **Oecologia**, v. 153, n. 3, p. 511-520, 2007.

HERRERA, C. M. Microclimate and individual variation in pollinators: Flowering plants are more than their flowers. **Ecology**, v.76, p. 1516– 1524, 1995.

HERRERA, C. M. Flower to seedling consequences of different pollination regimes in an insect-pollinated shrub. **Ecology**, v. 81, n. 1, p. 15-29, 2000.

HILL, J. L.; CURRAN, P. J. Species composition in fragmented forests: Conservation

implications of changing forest area. **Applied Geography**, v. 21. p. 157-174, 2001.

HODGSON, P.; FRENCH, K.; MAJOR, R. E. Avian movement across abrupt ecological edges: differential responses to housing density in an urban matrix. **Landscape and Urban Planning**, v. 79, n. 4, p. 266-272, 2007.

JULES, E. S.; SHAHANI, P. A broader ecological context to habitat fragmentation: Why matrix habitat is more important than we thought? **Journal of Vegetation Science**, v. 26, p. 29–44, 2003.

KEARNS, C. A.; INOUE, D. W.; WASER, N. M. Endangered mutualisms: the conservation of plant– pollinator interactions. **Annual Review of Ecology and Systematics**, v. 29, p. 83–112, 1998.

KRAUSS, J.; STEFFAN-DEWENTER, I; TSCHARNTKE, T. How does landscape context contribute to effects of habitat fragmentation on diversity and population density of butterflies? **Journal of Biogeography**, v. 30, p. 889–900, 2003.

KURKI, S.; NIKULA, A.; HELLE, P.; LINDEN, H. Landscape fragmentation and forest composition effects on grouse breeding success in boreal forests. **Ecology**, v. 81, p. 1985–97, 2000.

LAURANCE, W. F. Theory meets reality: how habitat fragmentation research has transcended island biogeographic theory. **Biological Conservation**, v. 141, p.1731–1744, 2008.

LAURANCE, W. F.; LOVEJOY, T. E.; VASCONCELOS, H. L.; BRUNA, E.M.; DIDHAM, R. K.; STOUFFER, P. C.; GASCON, C.; BIERREGAARD, R. O.; LAURANCE, S. G.; SAMPAIO, E. Ecosystem decay of Amazonian forest fragments: a 22-year investigation. **Conservation Biology**, v.16, p.605-618, 2002.

LAURENCE F. W. F.; VASCONCELOS H. L. Consequências ecológicas da fragmentação florestal na Amazônia. **Oecologia Brasiliensis**, v. 13, p. 434-451, 2009.

LEVINS, R. Some demographic and genetic consequences of environmental heterogeneity for biological control. **Bulletin of Entomological Society America**, v.15, p. 237-240, 1969.

LORD, J. M.; NORTON, D. A. 1990. Scale and the spatial concept of fragmentation. **Conservation Biology**, v. 4, p. 197– 202, 1990.

MACARTHUR, R. H.; WILSON, E. O. **The theory of island biogeography**. Princeton: Princeton University Press, 1967.

MAHAN, C. G.; YAHNER, R. H. Effects of forest fragmentation on behaviour patterns in the eastern chipmunk (*Tamias striatus*). **Canadian Journal of Zoology**, v. 77, p. 1991–97, 1999.

MCGARIGAL, K.; MARKS, B. J. **Fragstats: Spatial pattern analysis program for quantifying landscape structure**. General Technical Report PNW-GTR-351. Portland, OR: Pacific Northwest Research Station, 1995.

MEDELLÍN, R. A.; EQUIHUA, M.; AMIN, M. Bat diversity and abundance as indicators of disturbance in Neotropical rainforest. **Conservation Biology**, v. 14, p. 1666-1675, 2000.

MESQUITA, R. C. G.; ICKES, K.; GANADE, G.; WILLIAMSON, B. G. Alternative successional pathways in the Amazon basin. **Journal of Ecology**, v. 89, p. 528-537, 2001.

METZGER, J. P. Estrutura da paisagem e fragmentação: análise bibliográfica. **Anais da Academia Brasileira de Ciência**, 71, 445– 463, 1999.

METZGER, J. P. O que é ecologia de paisagens? **Biota Neotropica**, v.1, n. 1/2, p.1-9, 2001.

METZGER, J. P. Como restaurar a conectividade de paisagens fragmentadas? In: KAGEYAMA, P.Y.; OLIVEIRA, R. E.; MORAES, L. F. D.; ENGEL, V. L. **Ecossistemas naturais**. São Paulo: Fepaf, 2003. p. 49-76.

MONTERO-CASTAÑO, A.; VILÀ, M. Impact of landscape alteration and invasions on pollinators: a meta-analysis. **Journal of Ecology**, v. 100, p. 884–893, 2012.

MURCIA, C. Edge effects in fragment forests: implications for conservation. **Trends in Ecology and Evolution**, v.10, p. 58–62. 1995.

MURPHY, H. T.; LOVETT-DOUST, J. Context and connectivity in plant metapopulations and landscape mosaics: does the matrix matter? **Oikos**, v. 105, p. 3-14, 2004.

MYERS, N.; MITTERMEIER, R. A.; MITTERMEIER, C. G.; FONSECA, G. A. B.; KENT, J. Biodiversity hotspots for conservation priorities. **Nature**, v. 403, n. 6772, p. 853-858, 2000.

NAVEH, Z. Landscape ecology as an emerging branch of human ecosystem science. **Advances in Ecological Research**, v. 12, p. 189-237, 1982.

OPDAM, P.; VAN APELDOORN, R.; SCHOTMAN, A.; KALKHOVEN, J. **Population responses to landscape fragmentation**, In: Vos, C. C.; OPDAM, P. eds. Landscape ecology of a stressed environment, London, Chapman; Hall, p. 147-171. 1993.

OSADA, N.; TAKEDA, H.; FURUKAWA, A.; AWANG, M. Ontogenetic changes in leaf phenology of a canopy species, *Elateriospermum tapos* (Euphorbiaceae), in a Malaysian rain forest. **Journal of Tropical Ecology**, 18, 91–105. 2002.

PICKET, S. T. A.; CARDENASSO, M. L. Landscape ecology: spatial heterogeneity in ecological systems. **Science**, v. 269, p. 331-334, 1995.

PIÑERO, D.; SARUKHAN, J.; ALBERDI, P. The costs of reproduction in a tropical palm, *Astrocaryum mexicanum*. **Journal of Ecology**, v.70, p. 473–481, 1982.

PIVELLO, V. R.; METZGER, J. P. Diagnóstico da pesquisa em ecologia de paisagens no Brasil (2000-2005). *Biota neotropica*, v.7, p. 21-29, 2007.

PIZO, M.A., ALMEIDA-NETO, M. Determinants of fruit removal in *Geonoma pauciflora*, an understory palm of neotropical forests. **Ecological Research**, v. 24, p. 1179–1186, 2009.

PREVEDELLO, J. A.; VIEIRA, M. V. Does the type of matrix matter? A quantitative review of the evidence. **Biodiversity and Conservation**, v.19, p.1205–1223, 2010.

QUESADA, M.; STONER, K. E.; ROSAS-GUERRERO, V.; PALACIOS-GUE- VARA, C.; LOBO, J. A. Effects of habitat disruption on the activity of nectarivorous bats (Chiroptera: Phyllostomidae) in a dry tropical forest: implications for the reproductive success of the neotropical tree *Ceiba grandifolia*. **Oecologia**, v. 135, p. 400–406, 2003.

QUINTERO, C.; MORALES, C. L.; AIZEN, M. A. Effects of anthropogenic habitat disturbance on local pollinator diversity and species turnover across a precipitation gradient. **Biodiversity and Conservation**, v. 19, n. 1, p. 257-274, 2010.

RAMBALDI, D. M.; OLIVEIRA, D. A. S. **Fragmentação de ecossistemas: causas, efeitos sobre a biodiversidade e recomendações de políticas públicas**. Brasília: MMA/SBF, 2005.

RAMOS, F. N.; SANTOS, F. A. M. Phenology of *Psychotria tenuinervis* (Rubiaceae) in Atlantic forest fragments. **Canadian Journal of Botany**, v. 83, p.1305-1316, 2005.

RIBEIRO, C. M.; METZGER, J. P.; MARTENSEN, A. C.; PONZONI, F. J.; HIROTA M. M. The Brazilian Atlantic Forest: How much is left, and how is the remaining forest distributed? Implications for conservation. **Biological Conservation**, v. 142, n. 6, p. 1141-1153, 2009.

RICKETTS, T. H. The matrix matters: effective isolation in fragmented landscapes. **American Naturalist**, v. 158, p. 87–99, 2001.

RUDEL, T. K.; HOROWITZ, T. K. **Tropical deforestation: small farmers and land clearing in the Ecuadorian Amazon**. New York: Columbia University Press, 1993.

SANAIOTTI, T. M.; MAGNUSSON, W. E. Effects of annual fires on the production of fleshy fruits eaten by birds in a Brazilian Amazonian savanna. **Journal of Tropical Ecology**, v. 11, p. 53-65, 1995.

SAUNDERS, D. A.; HOBBS, R. J.; MARGULES, C. R. Biological consequences of ecosystem fragmentation: a review. **Conservation Biology**, v. 5, p.18-32, 1991.

SCARIOT, A.; FREITAS, S. R.; NETO, E.M.; NASCIMENTO, M. T.; OLIVIERA, L. C.; SANAIOTTI, T.; SEVILHA, A. C.; VILLELA, D. M. Vegetação e flora. In: RAMBALDI, D. M.; OLIVEIRA, D. A. S. **Fragmentação de ecossistemas: causas, efeitos sobre a biodiversidade e recomendações de políticas públicas**. Brasília: MMA/SBF, 2005.

SIH A.; BALTUS, M. Patch size, pollinator behavior and pollinator limitation in catnip. **Ecology**, v. 68, p. 1679–1690, 1987.

SOARES, N. C. **Variação intra-específica na fenologia de espécies de sub-bosque de floresta atlântica e sua relação com variáveis microambientais**. Dissertação (Mestrado em Biologia Vegetal), Universidade Estadual Paulista, Rio Claro, SP, 2011.

STEFFAN-DEWENTER, I.; MÜNZENBERG, U.; BÜRGER, C.; THIES, C., TSCHARNTKE, T. Scale-dependent effects of landscape context on three pollinator guilds. **Ecology**, v. 83, p. 1421–1432, 2002.

STONER, K. E. Murciélagos nectarívoros y frugívoros del bosque caducifolio de la Reserva de la Biosfera Chamela– Cuixmala. In: NOGUERA, F. A.; VEGA, J. A.; GARCIA-ALDRETE, G.; QUESADA, V. **Historia natural del bosque caducifolia de Chamela**, pp.

379–395. México City, México: Instituto de Biología, Universidad Nacional Autónoma de México, 2002.

STOUFFER, P. C.; BIERREGAARD, R. O.; STRONG, C.; LOVEJOY, T. E. Long-term landscape change and bird abundance in Amazonian rainforest fragments. **Conservation Biology**, v. 20, p.1212-1223, 2006.

STOUFFER, P. C.; BIERREGAARD, R. O. Use of Amazonian forest fragments by understory insectivorous birds. **Ecology**, v. 76, p. 2429-2445, 1995.

TAYLOR, P. D. L.; FAHRIG, K.; MERRIAM, H. Connectivity is a vital element of landscape structure. **Oikos**, v. 86, p. 571-573, 1993.

THEOBALD, D. M. Exploring the functional connectivity of landscapes using landscape networks. In: ROOKS, K.; SANJAYAN, M. A. (eds.). **Connectivity Conservation**. Cambridge University Press, Cambridge, NY. 2006, p. 417-443.

TOCHER, M. A comunidade de anfíbios da Amazônia central: diferenças na composição específica entre a mata primária e pastagens. Pp. 219-232. In: MOUTINHO, P.; GASCON, C. **Floresta Amazônica: Dinâmica, Regeneração e Manejo**. Manaus: Instituto Nacional de Pesquisas da Amazônia, 1998. 373p.

TROPPEMAIR, H. Ecologia da Paisagem: uma retrospectiva. Anais do I Fórum de debates “Ecologia da paisagem e planejamento ambiental”, Rio Claro, Sociedade de Ecologia do Brasil, 2000.

TURNER, M. G.; GARDNER, R. H.; O’NEILL, R. V. **Landscape ecology in theory and practice: patterns and process**. New York: Springer, 2001.

TURNER, M. G. Landscape ecology: What is the state of the science? **Annual Review of Ecology, Evolution, and Systematics**, v. 36, p. 319-344, 2005.

TURNER, I. M.; CORLETT, R. T. The conservation value of small, isolated fragments of lowland tropical rainforest. **Trends in Ecology and Evolution**, v. 11, p. 330-333, 1996.

TURNER, M. G.; GARDNER, R. H.; O’NEILL, R. V. **Landscape ecology in theory and practice: patterns and process**. New York: Springer, 2001.

UMETSU, F. **Pequenos mamíferos em um mosaico de habitats remanescentes e antropogênicos: qualidade da matriz e conectividade em uma paisagem fragmentada da Mata Atlântica.** Dissertação (Mestrado em Ecologia), Universidade de São Paulo, São Paulo, SP, 2005.

VAMOSI, J.C., KNIGHT, T.M., STEETS, J.A., MAZER, S.J., BURD, M., ASHMAN, T.L. Pollination decays in biodiversity hotspots. **Proceedings of the National Academy of Sciences.** USA. v.103, p. 956-961, 2006.

VIANA, B. F.; BOSCOLO, D., NETO, E.M., LOPES, L.E., LOPES, A.V.; FERREIRA, P.A., PIGOZZO, C.M. ; PRIMO, L.M. How well do we understand landscape effects on pollinators and pollination services? **Journal of Pollination Ecology**, v.7, p. 31- 34, 2012.

WAGENIUS, S., LONSDORF, E., NEUHAUSER, C. Patch aging and the S-allee effect: breeding system effects on the demographic response of plants to habitat fragmentation. **The American Naturalist**, v. 169, p. 383–397, 2007.

WASER, N. M.; PRICE, M. V. Outcrossing distance effects in *Delphinium nelsonii*: pollen loads, pollen tubes, and seed set. **Ecology**, v. 72, p. 171–179, 1991.

WEATHERS, K. C.; CADENASSO, M. L.; PICKETT, S. T. A. Forest edges as nutrient and pollutant concentrators: potential synergisms between fragmentation, forest canopies, and the atmosphere. **Conservation Biology**, v. 15, p. 1506–1514, 2001.

WETHERED, R.; LAWES, M. J. Matrix effects on bird assemblages in fragmented Afromontane forests in South Africa. **Biological Conservation**, v. 114, p. 327–340, 2003.

WIENS, J.A.; CRAWFORD, C. S.; GOSZ, J. R. Boundary dynamics: a conceptual framework for studying landscape ecosystems. **Oikos**, v. 45, p. 421–427, 1985.

WIENS, J.A.; MOSS, M.R. *Studies in Landscape ecology: issues and perspectives in landscape ecology.* Cambridge University Press. Cambridge, M. A. 2005.

WILCOVE, D.S.; MCLELLAN, C.H.; DOBSON, A.P. Habitat fragmentation in the temperate zone. In: *Conservation Biology*, ed. SOULÉ, M. E.; SUNDERLAND, M.A., SINAUER, 1986. p. 237–56.

WRIGHT, S. J.; VAN SCHAIK, C. P. Light and the phenology of tropical trees. **The American Naturalist**, v.143, p. 193–199, 1994.

WU, J.; HOBBS, R. J. **Key topics in landscape ecology**. New York: Cambridge University Press, p. 315, 2007.



## **SEGUNDA PARTE**

**ARTIGO: Proportion of habitat, type of matrix and fragment shape influence the plant reproduction**

**Autores:** Laura Fornero Gomes Ferreira, Erica Hasui e Flavio Nunes Ramos

**Artigo redigido conforme normas da Revista Biological Conservation.**

1 **PROPORTION OF HABITAT, TYPE OF MATRIX AND FRAGMENT SHAPE**  
2 **INFLUENCE THE PLANT REPRODUCTION**

3 Laura Fornero Gomes Ferreira<sup>1</sup>, Erica Hasui<sup>1</sup>, Flavio Nunes Ramos<sup>1,2</sup>

4 <sup>1</sup>Laboratório de Ecologia de Fragmentos Florestais (ECOFRAG), Instituto de Ciências da Natureza,  
5 Universidade Federal de Alfenas – UNIFAL-MG, Rua Gabriel Monteiro da Silva, 700, Centro, CEP  
6 37130-000, Alfenas, MG, Brasil. Telephone: +00553532991447.

7 <sup>2</sup> Corresponding author: fnramos@gmail.com

8 E-mail addresses: [laurafornero6@gmail.com](mailto:laurafornero6@gmail.com), [ericahasui@yahoo.com](mailto:ericahasui@yahoo.com)  
9

10 **ABSTRACT**

11  
12 Habitat loss and changes in the landscape structure are effects that results of the forest  
13 fragmentation and can negative affect ecological processes, as plant reproduction and  
14 pollination. We hypothesized that the landscape structure parameters (fragment area,  
15 proximity, percentage of landscape vegetation cover - native forest, coffee, sugarcane  
16 and pasture matrix, and fragment shape) and structural characteristics of plant  
17 (height stem circumference, circumference at breast height, plant height and volume of  
18 crown) and its habitat characteristics (canopy height, vertical profile, distance to the  
19 nearest individual and distance to the nearest conspecific individual) affect the flower  
20 and fruit production and number of floral visits of forest plant. We sampled 160  
21 individuals of a forest plant, *Psychotria vellosiana*, in eight landscapes (1000m radius),  
22 centered each one in one fragment, with 20 individuals per fragment. And we applied a  
23 general linear model and Akaike information criterion to select the best models that  
24 explains relations among plant reproduction with i) landscape and patch parameters and  
25 ii) measurements of structural characteristics of the plants and its habitat. We found that  
26 the number of floral visits and flower and fruit production are affected by the matrix  
27 type. Furthermore the total number of floral visits is influenced by the percentage of

28 forest concomitantly with percentage of coffee matrix. And fruit production is affected  
29 by landscape complexity, size of individual plant and degree of aggregation of the plant  
30 population. From these findings, we may conclude that landscapes characteristics are  
31 important to explain the plant reproduction. The information generated may contribute  
32 to the theoretical gap in knowledge on the influence of matrix on plants and on  
33 ecological process. In addition, they can be used in recommendations to farmers,  
34 decision makers and public policies, i.e. to inform what type of matrix should be used in  
35 fragmented landscape to help to maintain or improve the fruit production of forest  
36 plants.

37

38 Keywords: fragmentation; forest cover; matrix; plant fitness; pollination, proximity

39

40

#### 41 HIGHLIGHTS

42

- 43 • The type of matrix influenced the plant reproduction.
- 44 • The number of floral visits and fruit production was influenced by landscape  
45 structure.
- 46 • Fruit production was influenced by plant size and plant population density.

47

#### 48 1. Introduction

49

50 Landscape Ecology is an area of knowledge whose focus is to understand the  
51 influence of environmental spatial heterogeneity on ecological processes, emphasizing

52 human actions on the environment or the spatial context on populations (Pivello and  
53 Metzger 2007; Turner 2005). The landscape is a heterogeneous unit composed by a  
54 mosaic of interactive units, and its structure is defined by the area, shape and spatial  
55 arrangement of these units (Forman and Godron, 1986; Metzger, 1999). The landscape  
56 structure interferes on species survival and its movement across the landscape (Burgess,  
57 1988; Soulé et al., 1992). Thus, anthropic processes such as habitat fragmentation that  
58 alter this structure can modify ecological processes (Metzger, 1999; Wiens 2005).

59 Plant reproduction and pollination can be influenced by the loss of habitat and  
60 modification of landscape structure parameters, i.e. isolation of patches, type of matrix  
61 (Aizen and Feinsinger, 1994a; Viana et al., 2012). These parameters have great  
62 influence on plants and pollinators survival and dispersal capacity of pollinators, once  
63 they affect resources availability and control the functional connectivity of the  
64 landscape (Viana et al., 2012). The habitat removal can induce the pollen limitation and  
65 consequently negative affect the plant breeding success (Aizen and Feinsinger, 1994a,  
66 1994b; Fahrig, 2003; Kurki et al. 2000, Vamosi et al., 2006). Because the habitat loss  
67 can reduce the pollinators biodiversity, richness (Steffan-Dewenter et al. 2002),  
68 population abundance and distribution (Gibbs 1998, Guthery et al. 2001), change the  
69 aspects of floral visitors behavior that affect foraging success rate (Mahan and Yahner,  
70 1999) and can decrease the compatible plants (smaller density of plants and flowers) in  
71 fragments (Aizen and Feinsinger, 1994a; Griffin and Eckert, 2003; Wagenius et al.,  
72 2007). Besides, the patch isolation reduces the biodiversity and richness of pollinators  
73 and plants (Rukke 2000, Virgós 2001), is related to movement of animals among  
74 fragments (Bender et al., 2003; Tischendorf et al., 2003) and concomitantly with the  
75 type of matrix can be responsible by the habitat connectivity (Metzger, 1999). Thus, as

76 greater is the degree of isolation and resistance level of the surrounding matrix, lower  
77 the interpatch movement of pollinators is, minor the gene flow and the reproductive  
78 success of plant (Jules and Shahani, 2003; Viana et al. 2012).

79 Although the connectivity and structural characteristics of the mosaic landscape  
80 are important parameters that may interfere in ecological process (Aguilar et al., 2006;  
81 Aizen et al., 2002) still there is a lack of knowledge in this context (Ferreira et al., 2013;  
82 Prevedello and Vieira, 2010). Few studies relate the plant reproduction to landscape  
83 characteristics, especially to the type of matrix (Viana et al., 2012). Considering that the  
84 studied area (region on Atlantic Forest) is highly fragmented and has amount of habitat  
85 in extremely critical levels, understand what is happening to the species that are  
86 surviving in this environment becomes important to addressing the knowledge gap and  
87 to generate information that will help in matters of conservation and restoration  
88 (Ferreira et al., 2013; Morreale and Sullivan, 2010; Ribeiro et al., 2009).

89 The aim of this research was to investigate whether landscape structural  
90 parameters (landscape scale) and characteristics of plants and their habitat (local scale)  
91 influence the reproduction of forest plants, using as a model the shrub *Psychotria*  
92 *vellosiana*. We hypothesized that these factors mentioned above have a positive effect  
93 on plant reproduction, as they: (i) increase landscape vegetation cover of native forest  
94 (Lindenmayer and Luck, 2005), (ii) and proximity fragment (Farwig et al., 2009), (iii)  
95 matrices more similar to forest environment (coffee) (Prevedello and Vieira, 2010), (iv)  
96 patches with more regular form (McGarigal and Marks, 1995; Murcia, 1995), (v) plants  
97 with greater values of structural measurements, such as CAS, CAP, plant height and  
98 ellipsoid volume of and (vi) variations in values of measurements of traits of plant  
99 habitat, such as canopy height, vertical profile of the understory, distance to the nearest

100 individual and distance to the nearest conspecific individual (Blendinger et al., 2008;  
101 Pizo and Almeida-Neto, 2009). We found that the total number of floral visits, likewise  
102 the flower and fruit production, are influenced by matrix type. Moreover, the number of  
103 floral visits is affected by the percentage of forest cover in the landscape combined with  
104 the percentage of coffee matrix, and fruit production by landscape complexity, size of  
105 individual plant and degree of aggregation of the plant population.

106

## 107 2. Methods

108

### 109 2.1. Study Area and Landscape Selection

110

111 The study area is into domain of Brazilian Atlantic Forest, in region of Alfenas  
112 city (84722.62 ha) (Table 1), south of Minas Gerais, Brazil (Fundação SOS Mata  
113 Atlântica, 2008). The status of conservation of this area tends to critical due to the  
114 advancement of habitat loss and fragmentation, the percentage of habitat in this region  
115 is represented by 4% of forest fragments (Fundação SOS Mata Atlântica, 2008), 97 %  
116 of the fragments are < 50 ha, the median of ENN (Euclidean nearest neighbor distance,  
117 used to quantify the isolation of fragments habitat) is 212 m (standard deviation= 301)  
118 and the largest fragment has 203 ha (unpublished data). The economy in the region is  
119 mainly based on agriculture, mainly coffee plantations (13500 ha) and others as  
120 sugarcane, corn and pasture (IBGE, 2006). The vegetation of this study was classified  
121 as Semideciduos Atlantic forest and the climate was classified as Cwa according to the  
122 Köppen system (Vianello and Alves, 1991) since the summer is hot and warm, and the  
123 winter is dry.

124 We selected the fragments from the classification of remnant forest areas around  
125 Alfenas city, and resort digital processing of satellite images of Sino-Brasileiro CBERS-  
126 2B, of 2009, with a resolution of 20 meters. The attributes used to select the fragments  
127 were: i) remnants with similarity at the degradation level, based on image texture ii)  
128 presence of at least one of three types of matrix (shrubby perennial represented by  
129 coffee, semi-perennial grass by sugarcane and perennial grass, by pasture) and iii)  
130 presence of the species model, *Psychotria vellosiana*. We done eight landscapes (1000m  
131 radius) each one with 314 ha area and centered in one forest fragment selected. We  
132 measured landscape and fragment metrics for each of the eight central fragments. And  
133 to analyze plant reproduction (flower and fruit production, floral visits, fruit and seed  
134 set, plants and its microenvironment traits) we sampled 20 adult individuals of  
135 *Psychotria vellosiana*, higher than 1.5m, per transects with until 40m (one per  
136 fragment), in the interior (environment as close as possible to the center of fragment) of  
137 each fragment (total 160 individuals).

138

## 139 2.2. *Psychotria vellosiana* as Species Model

140

141 Rubiaceae family, 400-500 genera (Barroso, 1991), is composed of trees and  
142 mainly shrubs (Cronquist, 1981). The genus *Psychotria* is the largest in the family  
143 Rubiaceae (Davis et al., 2001), having distylous flowers (pin and thrum), with the  
144 majority of species showing incompatibility within the same individual and morph  
145 (Bawa and Beach, 1983; Hamilton, 1990). Most species of this genus show elevated  
146 synchrony within a population, flowering during the rainy season, with the fruit  
147 remaining unripe for some months (Almeida and Alves, 2000; Castro and Araújo, 2004;

148 Ramos and Santos, 2005; Virillo et al., 2007). *Psychotria vellosiana* Benth. is a shrub or  
149 small tree, 2.5 to 4 m tall, usually occurring in shaded sites and distributed in eastern  
150 Brazil (Araújo and Cardoso, 2006). The flowers are small and pollinated by insects,  
151 mainly bees like other *Psychotria* species (Castro and Araújo, 2004; Hamilton, 1990).

152         The shrub was chosen because it is a good model to study the effect of  
153 fragmentation on plant reproduction (Lopes and Buzato, 2007). It can be considered a  
154 good model because is abundant in the fragments studied, produced flower and fruits at  
155 a relatively low height above the ground, must be an exigent species in relation to  
156 temperature and water availability, since most species of *Psychotria* are common in the  
157 understory forest (Hamilton, 1990; Mabberley, 1997; Taylor, 1996) and shaded areas  
158 (Pietrobon et al., 2011), but not so exigent that does not respond to changes caused by  
159 fragmentation.

160

### 161 2.3. Quantification of landscape structure

162

163         The landscape analysis was done with an ArcGIS 10.0 (Rempel, 2010). We  
164 made buffers of 1000m around each fragment chosen and the total area inside the buffer  
165 was classified. We choose this size of buffer because it is believed that various species  
166 of bees do not forage great distances; Gathmann and Tschardtke (2002) studied 16 wild  
167 bee species and investigated foraging distances between nesting sites and suitable food  
168 plants and found maximum flight distances ranging from 150 to 600m. We did the  
169 manual classification of the image to identify the forest fragments, calculate the  
170 percentage of vegetation cover, percentage of matrix (discriminate types of use) and  
171 evaluate the area structures. We classified the polygons as forest, coffee, sugarcane and



172 pasture percentage. Furthermore, we calculated the landscape metrics based on vector  
173 map by vLATE (Lang and Tiede, 2003) and the Patch Analyst Extension of ArcGIS  
174 10.0 (Rempel, 2010): i) Fragment area (ha), ii) PROX (m) (proximity fragment is a  
175 dimensionless index inversely related to the isolation of the fragment, where within a  
176 radius of 1 km, it is considered both distance and the area of the fragments present, as  
177 the larger the fragment neighbors are and the closer they are to the target fragment, the  
178 greater the value of proximity is), iii) Percentage of landscape vegetation cover  
179 (percentage of the class - native forest, coffee, sugarcane and pasture matrix calculated  
180 from the class area) and iv) Area Weighted Mean Shape Index (calculate the irregularity  
181 of remaining landscape performing a weighted average for the area. The index values  
182 are equal to 1 when all patches are circular, and it increases with increasing patch shape  
183 irregularity and area) (McGarigal and Marks, 1995). In order to check if these metrics  
184 influence the variables related to plant reproduction, we did various assessments.

185

#### 186 2.4. Plant and habitat characteristics

187

188 For each individual plant sampled (160 total individuals), we measured  
189 structural and habitat characteristics. The structural plant measurements were: height at  
190 soil circumference (CSH), circumference at breast height (CBH), plant height, ellipsoid  
191 volume of crown ( $\frac{4}{3}\pi abc$ ,  $\pi$  is a constant equal 3.14, **a** and **b** are the major and minor  
192 diameters of crown and **c** is the height crown or total height of plant minus the height of  
193 the fork of shrub) (Zotz et al., 1999). Plant habitat characteristics (related to light  
194 availability or floral visits) measured were: canopy height just above each individual  
195 plant (m), vertical profile of the understory (or vertical density, counting the number of

196 times the vegetation touched the pole, 6.0 m height, suspended vertically at breast  
197 height to above the crown of each individual) according to the methodology suggested  
198 by Pizo and Almeida-Neto (2009), distance to the nearest individual and distance to the  
199 nearest conspecific individual of each *Psychotria vellosiana* sampled. These  
200 measurements were taken to verify whether plant traits and their microenvironment  
201 features may influence plant reproduction.

202

### 203 2.5. Phenology

204

205 We did monthly phenological observations for one year, between July (2011)  
206 and June (2012), of 160 individuals of *P. vellosiana*, inserted in eight forest fragments.  
207 In each fragment we sampled twenty adult individuals higher than 1.5 m, along of one  
208 transect until 40m, per fragment, in the center of the remnant. To estimate phenological  
209 patterns of plants, we applied the Fournier intensity method (Fournier, 1974). The  
210 Fournier index estimates the intensity of each phenophase through a semi-quantitative  
211 scale, where 0 = absence of characteristic; 1 = presence of the characteristic with  
212 intensity from 1 to 25%, 2 = presence of the characteristic with intensity of > 25 to 50%,  
213 3 = presence feature with an intensity of > 50 to 75%, 4 = presence of the characteristic  
214 intensity greater than 75% (Fournier, 1974). To estimate flower and fruit production  
215 (ripe fruit), we calculated the proportion of annual intensity per individual, for local  
216 scale analyses and the proportion of annual intensity per fragment, for landscape scale  
217 analyses. For the local scale, we found the sum of intensity values obtained for each  
218 individual for 12 months, divided by the number of individuals multiplied by four  
219 (maximum value of intensity that the plant may present) and by the number of months.

220 The proportion obtained was multiplied by 100 to convert it to a percentage value. For  
221 the landscape scale, we found the sum of intensity values obtained for the 20 individuals  
222 of a fragment for 12 months, divided by the same formula presented previously and  
223 multiplied by 100 (Fournier, 1974).

224

## 225 2.6. Floral visits

226

227 We accompanied a total of 32 individuals of *P. vellosiana* to obtain data of floral  
228 visits. For the purpose of minimizing the possibility that the number of flowers could  
229 influence the attraction of visitors, we sampled four individuals (2 pin and 2 thrum) with  
230 a similar number of flowers (25–50% of the crown with flowers, which corresponded to  
231 the most frequently flowering individuals in the area) along a transect of 40 meters in  
232 the interior of each fragment. The observations were made between 0700-1700 h. We  
233 observed each shrub (restricted to a branch or part of the crown) for 10 min in three  
234 different days (for total of 96 period of 10 min), and during this period we recorded the  
235 number of species of diurnal floral visitors and the frequency of visits (sensu Aizen and  
236 Feinsinger, 1994b). The visits were recorded in November and December in 2011  
237 (approximately 15 days) when plants bloomed; no individual of *P. vellosiana* was  
238 observed twice during the same hour of the day in order to avoid systematic bias  
239 introduced by changes in insect behavior (sensu Ramos and Santos, 2006). To obtain  
240 the total value of number of species and frequency of floral visits per fragment, we did a  
241 sum of all of them in each fragment (four sampled plants). These observations were  
242 made to verify whether landscape configuration may influence floral visits, important  
243 processes related to plant reproduction. We collected at least one individual for each

244 visiting species, it was identified by a specialist and deposited in the Instruction  
245 Laboratory of the Zoology of Federal University of Alfenas (Unifal-MG).

246

247 2.7. Fruit and seed set

248

249 To analyze whether the landscape affects the proportion of fruit and seed  
250 developed, we quantified fruit production per inflorescence counting the ripe fruit  
251 divided by the number of flowers originally produced (total number of scars) in six  
252 infructescences randomly sampled from five shrubs in each landscape (*sensu* Ramos  
253 and Santos, 2006). And to estimate the seed set, we counted the number of developed  
254 seeds of 15 fruits per shrub randomly sampled (five shrubs per landscape, total of 40  
255 shrubs).

256

257 2.8. Data analysis

258

259 2.8.1. Relation of flower and fruit production with plant traits and their habitats

260

261 We tested the dependent variables, data of individual plant production (flower  
262 and ripe fruit), as well as independent variables, plant traits (CSH, CBH, plant height,  
263 ellipsoid volume of crown) and extrinsic traits (canopy height above the individual,  
264 understory vertical density, distance to the nearest individual and distance to the nearest  
265 conspecific individual) to verify whether they show normal distribution (D'Agostino  
266 test). The analysis of the relationships of independent variables with the dependent  
267 variables was done by constructing and selecting General linear models (GLM). The

268 models constructed include the effects of one variable or two independent variable  
269 combined (not correlated, Spearman correlation coefficient), to represent current  
270 hypotheses in the literature on the effects of plant traits and extrinsic traits on the flower  
271 and fruit production (Table 2 and 3). We used the Akaike information criterion (AIC) to  
272 select the best model (general linear model) (Burnham and Anderson, 2002).

273         The AIC criterion considers the principle of parsimony, because it considers the  
274 model that fit the data as well as penalizing model complexity or number of variables  
275 (Bolker, 2006). The models with the lowest values of AIC are the most plausible. The  
276 correction AICc recommended to adjust models in relatively small samples was  
277 performed to calculate the AIC of our models. To compare the plausibility of the  
278 models, the  $\Delta AICc$  and the relative weight of each model ( $w$ ) was calculated. Models  
279 with  $\Delta AICc \leq 2$  and  $w \geq 0.10$  were considered equally plausible to explain the data  
280 (Burnham and Anderson, 2002).

281

#### 282 2.8.2. Relation of plant reproduction parameters with landscape metrics

283

284         The dependent variables were represented by plant population production (flower  
285 and ripe fruit), fruit set, seed set and floral visits (number of species of floral visitors  
286 and the frequency of visits). And the independent variables were represented by  
287 landscape and patch metrics (Fragment area, PROX, Percentage of plant coverage  
288 (native forest and matrix of coffee, sugarcane and pasture) and Area weighted mean  
289 shape index). Were tested these variables to verify whether they show normal  
290 distribution (Shapiro-Wilk test). We did the analysis of the relationships of variables  
291 and the construction of models as mentioned in the previous section. In addition to

292 Spearman, we did a Pearson linear correlation to exclude the correlated variables. The  
293 models represent hypotheses about landscape and patch metrics effect on plant  
294 reproduction (Table 3 and 4). And we selected the best model according AIC (Burnham  
295 and Anderson, 2002).

296

### 297 3. Results

298

299 Both scales influenced plant reproduction. In the landscape scale, the parameter  
300 matrix type, area weighted mean shape index (irregularity of remaining landscape) and  
301 percentage of forest cover influence some reproductive aspects of plant. The total  
302 number of floral visits was positively influenced by the percentage of native forest  
303 associated with percentage of coffee matrix (Table 5, Fig. 3). Among the five most  
304 abundant species of floral visitors (Table 6), four were influenced by parameters of  
305 landscape structure. The number of visits of *Partamona cf. cupira*, the most frequent  
306 flower visitor, was positively influenced by the percentage of forest cover, as well the  
307 fourth most abundant species, *Apis mellifera*, which was also positively influenced by  
308 the increase of percentage of sugarcane and pasture matrix. The second most abundant  
309 floral visitor, Hymenoptera sp, in contrast was negatively affected by the increase of  
310 percentage of forest cover. And the third more abundant species, *Paratetrapedia fervida*  
311 was positively influenced by the increase of coffee and sugarcane matrix (Table 5).

312 Four fragments (50%) presented more than 30% of their landscape covered by  
313 forest and coffee plantations, and the coffee was crucial to achieve or surpass this  
314 threshold (Fig. 2). The increasing of percentage of pasture and sugarcane matrix  
315 enhanced the flower production and the decreasing of: i) percentage of coffee matrix

316 and ii) the irregularity of remaining landscape had a positive effect on fruit production  
317 (Table 5).

318 Additionally, on the local scale, the parameter plant height and distance to the  
319 nearest plant conspecific individual influenced reproductive aspects. The increase of  
320 plant height and the decrease of distance of nearest conspecific plant had a positive  
321 effect on fruit production (Table 5).

322

#### 323 4. Discussion

324

325 Our findings demonstrate that flower and fruit production, as well as the number  
326 of floral visits are influenced by the matrix type. Additionally the number of floral visits  
327 is influenced by percentage of forest, associated with percentage of coffee matrix, and  
328 fruit production by landscape complexity, size of individual plant and degree of  
329 aggregation of the plant population.

330 In the present study, the increase of percentage of pasture and sugarcane matrix  
331 around forest fragments had a positive effect on flower production. Annual grass  
332 cultures, compared with other matrices, such as the perennial shrub matrix, are less  
333 similar to forest environment, due to their vegetal stratification and height variation  
334 throughout the year (Jules and Shahani, 2003; Malcolm, 1991). Therefore, pasture and  
335 sugarcane matrix, contrary to forest, allowed greater solar radiation to reach the ground  
336 during the day and irradiation to the atmosphere during the night (Murcia, 1995).  
337 Consequently the temperature in these matrices tends to reach higher maximum  
338 (Murcia, 1995). Besides, these matrices can permit a large input of direct or indirect  
339 light incidence and increase of temperature on the remnant: factors that positively

340 influence the flowering phase, increasing the flower production on plant species (Wright  
341 and van Schaik, 1994).

342           Landscape vegetation covered with native forest, associated with shrubby  
343 perennial matrix (in the present study represented by coffee) positively influences the  
344 number of floral visits in *P. vellosiana*. First, we can infer that forest cover is important  
345 in maintaining floral visitors in the landscape (Fortuna and Bascompte, 2006; Ricketts,  
346 2004), because it can provide suitable biotic (vegetal stratification) and abiotic  
347 (microclimate) conditions for insect nesting, foraging and dislocation (Rambaldi and  
348 Oliveira, 2005). Second, the coffee matrix may maintain or help maintain the forest  
349 function for pollinators (Greenberg et al., 1997; Moguel and Toledo, 1999; Ricketts,  
350 2004). Since this matrix type belongs to the same family, Rubiaceae, of the specie  
351 model *P. vellosiana* and because presented intermediate-sized plants (2 - 4m plant adult  
352 height) and is perennial. These last characteristics improved the microclimate  
353 (irradiation, air and soil humidity, wind velocity, etc.), being more similar to a forest  
354 environment, contrasted to pasture or other annual cultures (Renjifo, 2001). Hence,  
355 connectivity can be improved among patches, in a landscape combining a forest and  
356 perennial shrub matrix, facilitating insect dislocation, necessary for pollen flow (Muriel  
357 and Kattan, 2009; Ricketts, 2004). Furthermore, coffee, and *P. vellosiana* belong to the  
358 same taxonomic family, Rubiaceae, having common floral blooming in the same period.  
359 Coffee blooming between September - December in Brazil (Camargo and Camargo,  
360 2001) and October - November in south of Minas Gerais state, according to  
361 ANDRADE, H. C. C. (consultant on issues related to coffee) (pers. comm., November  
362 2012). In turn, *Psychotria vellosiana* blooms between November - December in the  
363 studied region (Personal observation). This overlap would be an advantage for the



364 reproduction of *P. vellosiana*, since a coffee matrix may attract and maintain flower  
365 visitor insects in the landscape (Jules and Shahani, 2003).

366         One important contribution of the perennial shrub matrix was to help achieve the  
367 ecological threshold in the landscape (30% of native vegetation) when added to a  
368 percentage of native forest. The ecological threshold can be defined as a zone in which  
369 rapid changes occur in ecological conditions (Bennett and Radford, 2003), thus whether  
370 native forest coverage becomes less than 30% (Andr n, 1994; Swift and Hannon, 2010),  
371 there is a sudden drop in the number of original species due to the higher intensity of  
372 fragmentation effects (Lindenmayer and Luck, 2005; Metzger and D camps, 1997).  
373 Consequently, the perennial shrub matrix contribution would facilitate the maintenance  
374 of pollinators in the region. Thus, in this study, the two habitats (native forest and coffee  
375 matrix) together may improve the connectivity and the persistence of populations of  
376 flower visitors in a fragmented landscape. In addition, *Partamona cf. cupira*, the highest  
377 frequency flower visitor can be considered a forest specialist, since its number of floral  
378 visits was positively affected by the increase of native forest, in the present study  
379 (Murcia, 1996). The other most frequent species may be considered generalists or open  
380 area species, because its visits were positively affected by various types of habitats and  
381 with different degrees of disturbance (Table 5) (Murcia, 1996).

382         We also found that, fruit production is positively influenced by the decrease of  
383 percentage of coffee matrix and irregularity of patch shape. Although, the increase of  
384 percentage of coffee matrix and native forest have favored the increase in number of  
385 floral visits (pollen quantity), it was not sufficient to increase fruit production.  
386 Reproductive success also depends on the origin of deposited pollen (pollen quality)  
387 (Aizen and Harder, 2007; Waser and Price, 1991) and, thus, plants that receive high

388 rates of pollen of related individuals have a reduced fruit and seed set (Charlesworth and  
389 Charlesworth, 1987; Griffin and Eckert, 2003; Herlihy and Eckert, 2004). As *P.*  
390 *vellosiana* shows incompatibility within the same individual and morph (Hamilton,  
391 1990) the chance of reproductive success in fragmented areas may be lower. Moreover,  
392 as the most frequent floral visitor may be a forest specialist, and probably they cannot  
393 easily cross between patches, enhancing inbreeding, reducing gene flow and increasing  
394 the chance of exchanging pollen between related plants (Shapcott, 1998). One study,  
395 that investigated the influence of habitat complexity and landscape configuration on  
396 pollination of wild cherry trees, also did not find a relationship between pollinator  
397 visitation rates and fruit formation, due to the differences in bee foraging and therefore  
398 in the quality of deposited pollen (Breitbach et al., 2012). Additionally, we can  
399 hypothesize that whether a coffee matrix increases the quantity of floral visits, but  
400 decreases fruit production, the floral visitors favored by coffee may not necessarily be  
401 the effective pollinators of *P. vellosiana*.

402         Furthermore, we found that fruit production is positively affected by a variation  
403 in patch shape. In other words, the more regular and similar to a circle the shapes of the  
404 patches are, the greater the fruit production is (McGarigal and Marks, 1995; Murcia,  
405 1995). Fragments with a more regular shape have less edge effect (McGarigal and  
406 Marks, 1995), and as the fruit formation depends on the performance of the parental  
407 plant, any change in abiotic factors (such as hydric stress, change in nutrient  
408 availability, high temperature, herbivory) in fragments can negatively affect fruit  
409 production (Ghazoul, 2005; Knight et al., 2005; Wesselingh, 2007).

410         Finally, we found that forest specialist plants with larger size and near  
411 conspecific (population with high aggregation) individuals should produce more fruit.

412 This last relation, between proximity among conspecifics and increase of fruit  
413 production had not yet been reported in the literature. While smaller plants invest their  
414 resources in the production of root and leaf system, larger plants, closer to the adult  
415 phase, allocate their resources in reproduction. Thus larger plants tend to produce more  
416 flowers, attract more pollinators and consequently produce more fruit (Piñero et al.,  
417 1982; Osada et al., 2002). The short distance between conspecifics, associated with the  
418 fact that *Psychotria vellosiana* is an entomophilous species and usually aggregate in  
419 flowering activity (Morellato, 1991; Smith-Ramirez and Armesto, 1994) may improve  
420 the pollinator attraction and the chance of pollination and fruit production (Blendinger  
421 et al., 2008; Ramos and Santos, 2005). Moreover, the presence of neighboring fruiting  
422 conspecifics increases the visibility and attractiveness, and it may increase the visits of  
423 seed dispersers and plant fitness (Blendinger et al., 2008; Carlo and Morales, 2008).

424

## 425 5. Conclusion

426

427 From our results, we can conclude that the production of flower and fruit and the  
428 total number of floral visits are affected by the matrix type. Furthermore, the number of  
429 floral visits is influenced by percentage of forest cover combined with the percentage of  
430 coffee matrix, and fruit production by landscape complexity, size of individual plant  
431 and degree of aggregation of the plant population. From these findings we can affirm  
432 that landscapes characteristics are important to plant reproduction and we may  
433 contribute to the theoretical gap in knowledge about the influence of matrix on  
434 ecological process and mainly on plant (Ferreira et al., 2013). Furthermore, information  
435 generated by this study can be used in recommendations to farmers, decision makers

436 and public policies, i.e. what type of matrix should be used to enhance the landscape  
437 connectivity and consequently help to maintain or improve the fruit production of forest  
438 plants in fragmented areas.

439

#### 440 Acknowledgements

441 The authors thank Antonio Jose Camillo de Aguiar for identifying the floral visitors,  
442 Gabrielle Carnevalli Vilela, Marcela Albertini Roquim Alcantara and André Araújo  
443 Carvalho for help with the fieldwork. CAPES, CNPq, FAPEMIG and Vale Company  
444 for the financial support. Jim Hesson of AcademicEnglishSolutions.com corrected the  
445 English.

446

#### 447 REFERENCES

448

- 449 Aguilar, R., Ashworth, L., Galetto, L., Aizen, M.A., 2006. Plant reproductive  
450 susceptibility to habitat fragmentation: review and synthesis through a meta-  
451 analysis. *Ecol. Lett.* 9, 968–980.
- 452 Aizen, M.A., Feinsinger, P., 1994a. Forest fragmentation, pollination, and plant  
453 reproduction in a Chaco dry forest, Argentina. *Ecology* 75, 33–351.
- 454 Aizen, M.A., Feinsinger, P., 1994b. Habitat fragmentation, native insect pollinator, and  
455 feral honey bees in Argentina “Chaco Serrano”. *Ecol. Appl.* 4, 378–392.
- 456 Aizen, M.A., Harder, L.D., 2007. Expanding the limits of the pollen limitation concept:  
457 effects of pollen quantity and quality. *Ecology* 88, 271–281.

- 458 Aizen, M.A., Ashworth, L., Galetto, L., 2002. Reproductive success in fragmented  
459 habitats: do compatibility systems and pollination specialization matter? *J. Veg.*  
460 *Sci.* 13, 885–892.
- 461 Almeida, E.M., Alves, M.A.S., 2000. Fenologia de *Psychotria nuda* e *P. brasiliensis*  
462 (Rubiaceae) em uma área de floresta Atlântica no sudeste do Brasil. *Acta Bot. Bras.*  
463 14, 335–346.
- 464 Andrén, H., 1994. Effects of habitat fragmentation on birds and mammals in landscapes  
465 with different proportions of suitable habitat: a review. *Oikos* 71, 355–366.
- 466 Araújo, C.G., Cardoso, V.J.M., 2006. Storage in cerrado soil and germination of  
467 *Psychotria vellosiana* (Rubiaceae) seeds. *Braz. J. Biol.* 66, 709–717.
- 468 Barroso, G.M., 1991. Sistemática de Angiospermas do Brasil. Imprensa Universitária,  
469 Universidade Federal de Viçosa, Viçosa.
- 470 Bawa, K.S., Beach, J.H., 1983. Self-incompatibility systems in the Rubiaceae of a  
471 tropical lowland wet forest. *Am. J. Bot.* 70, 1281–1288.
- 472 Bender, D.J., Tischendorf, L., Fahrig L., 2003. Evaluation of patch isolation metrics for  
473 predicting animal movement in binary landscapes. *Landsc. Ecol.* 18, 17–39
- 474 Bennett, A., Radford, J., 2003. Know your ecological thresholds. *Thinking Bush* 2, 1–  
475 3.
- 476 Blendinger, P.G., Loiselle, B.A., Blake, J.G., 2008. Crop size, plant aggregation, and  
477 microhabitat type affect fruit removal by birds from individual melastome plants in  
478 the Upper Amazon. *Oecologia* 158, 273–283.
- 479 Bolker, B., 2006. *Ecological Models and Data in R*. Princeton University Press, New  
480 Jersey.

- 481 Breitbach, N., Tillmann, S., Schleuning, M., Grunewald C., Laube, I., Steffan-  
482 Dewenter, I. Böhning-Gaese, K., 2012. Influence of habitat complexity and  
483 landscape configuration on pollination and seed-dispersal interactions of wild cherry  
484 trees. *Oecologia* 168, 425–437.
- 485 Burgess, V.J., Kelly, D., Robertson, A.W., Ladley, J. J., 2006. Positive effects of forest  
486 edges on plant reproduction: literature review and a case study of bee visitation to  
487 flowers of *Peraxilla tetrapetala* (Loranthaceae). *New Zea. J. Ecol.* 30, 1–13.
- 488 Burnham, K.P., Anderson, D.R., 2002. Model selection and multimodel inference: a  
489 practical information-theoretic approach, Second ed. Springer- Verlag, New York.
- 490 Camargo, A.P., Camargo, M.B.P., 2001. Definição e esquematização das fases  
491 fenológicas do cafeeiro arábica nas condições tropicais do Brasil. *Bragantia* 60, 65–  
492 68.
- 493 Carlo, T.A., Morales, J.M., 2008. Inequalities in fruit-removal and seed dispersal:  
494 consequences of bird behaviour, neighbourhood density and landscape aggregation.  
495 *J. Ecol.* 96, 609–618.
- 496 Castro, C.C., Araújo, A.C., 2004. Distyly and sequential pollinators of *Psychotria nuda*  
497 (Rubiaceae) in the Atlantic rain forest, Brazil. *Plant Syst. Evol.* 244, 131–139.
- 498 Charlesworth, B., Leon, J.A., 1976. The relation of reproductive effort to age. *Am. Nat.*  
499 110, 449–459.
- 500 Charlesworth, D., Charlesworth, B., 1987. Inbreeding depression and its evolutionary  
501 consequences. *Annu. Rev. Ecol. Syst.* 18, 237–268.
- 502 Cronquist, A., 1981. An integrated system of classification of flowering plants.  
503 Columbia University Press, New York.

- 504 Davis, A.P., Bridson, D., Jarvis, C., Govaerts, R., 2001. The typification and  
505 characterization of genus *Psychotria* L. (Rubiaceae). *Bot. J. Linn. Soc.* 135, 35–42.
- 506 Duffy, K.J., Stout, J.C., 2011. Effects of conspecific and heterospecific floral density on  
507 the pollination of two related rewarding Orchids. *Plant. Ecol.* 212, 397–1406.
- 508 Fahrig, L., 2003. Effects of habitat fragmentation on biodiversity. *Annu. Rev. Ecol.*  
509 *Evol. Syst.* 34, 487–515.
- 510 Farwig, N., Bailey, D., Bochud, E., Herrmann, J.D., Kindler, E., Reusser, N., C.  
511 Schüepp., Schmidt-Entling M.H., 2009. Isolation from forest reduces pollination,  
512 seed predation and insect scavenging in Swiss farmland. *Landscape Ecol.* 24, 919–  
513 927.
- 514 Ferreira, P.A., Boscolo, B., Blandina, F.V., 2013. What do we know about the effects of  
515 landscape changes on plant–pollinator interaction networks? *Ecol. Indicat.*, in press.  
516 Available from <http://dx.doi.org/10.1016/j.ecolind.2012.07.025> (accessed on 9  
517 January 2012).
- 518 Forman, R.T.T., Godron, M., 1986. *Landscape ecology*. Wiley and Sons Ed., New  
519 York.
- 520 Fortuna, M.A., Bascompte, J., 2006. Habitat loss and the structure of plant–animal  
521 mutualistic networks. *Ecol. Lett.* 9, 281–286.
- 522 Fournier, L.A., 1974. Um método quantitativo para la medición de características  
523 fragmentation: a synthesis. *Global Ecol. Biogeogr.* 16, 265–280.
- 524 Fundação SOS Mata Atlântica, 2008. **Atlas dos remanescentes florestais de Mata**  
525 **Atlântica - Período 2000 a 2005**. São Paulo: SOS Mata Atlântica/INPE/ISA.
- 526 Gathmann, A., Tschardt, T., 2002. Foraging distances of solitary bees. *J. Anim. Ecol.*  
527 71, 757–764.

- 528 Ghazoul, J., 2005. Pollen and seed dispersal among dispersed plants. *Biol. Rev.* 80,  
529 413–443.
- 530 Gibbs, J.P., 1998. Distribution of woodland amphibians along a forest fragmentation  
531 gradient. *Landscape Ecol.* 13, 263–68.
- 532 Greenberg, R., Bichier, P., Cruz, A. Reitsma, R., 1997. Bird populations in shade and  
533 sun-coffee plantations in central Guatemala. *Conserv. Biol.* 11, 448–459.
- 534 Griffin, C.A.M., Eckert, C.G., 2003. Experimental analysis of biparental inbreeding in a  
535 self-fertilizing plant. *Evolution* 57, 1513–1519.
- 536 Guthery, F.S., Green M.C., Masters R.E., Demaso, S. J., Wilson, H.M., Steubing, F.B.,  
537 2001. **Landcover and bobwhite abundance on Oklahoma farms and ranches.** *J.*  
538 *Wildl. Manage.* 65, 838–49.
- 539 Hamilton, C.W., 1990. Variations in a distylous theme in mesoamerican *Psychotria*  
540 subgenus *Psychotria* (Rubiaceae). *Mem. New York Bot. Gard.* 55, 62–75.
- 541 Herlihy, C.R., Eckert, C.G., 2004. Experimental dissection of inbreeding and its  
542 adaptive significance in a flowering plant, *Aquilegia canadensis* (Ranunculaceae).  
543 *Evolution* 58, 2693–2703.
- 544 IBGE. Instituto brasileiro de geografia e estatística. Produção agrícola municipal 2006.  
545 Disponível em: <[http://www.ibge.gov.br/home/presidencia /noticias/  
546 noticia\\_visualiza.php?id\\_noticia=998](http://www.ibge.gov.br/home/presidencia /noticias/ noticia_visualiza.php?id_noticia=998) >.accessed in 12 março 2013
- 547 Jules, E.S., Shahani, P., 2003. A broader ecological context to habitat fragmentation:  
548 Why matrix habitat is more important than we thought? *J. Veg. Sci.* 26, 29–44.
- 549 Kato, E., Hiura T., 1999. Fruit set in *Styrax obassia* (Styracaceae): the effect of light  
550 availability, display size, and local floral density. *Am. J. Bot.* 86, 495–501.



- 551 Knight, T.M., Steets, J.A., Vamosi, J.C., Mazer, S.J., Burd, M. Campbell, D.R., Dudash,  
552 M.R., Johnston, M.O., Mitchell, R.J., Ashman, T.L., 2005. Pollen limitation of plant  
553 reproduction: pattern and process. *Annu. Rev. Ecol. Evol. Syst.* 36, 467–497.
- 554 Kurki, S., Nikula, A., Helle, P., Linden, H., 2000. Landscape fragmentation and forest  
555 composition effects on grouse breeding success in boreal forests. *Ecology*, 81,  
556 1985–97.
- 557 Lang, S., Tiede, D., 2003. vLATE Extension für ArcGIS - vektorbasiertes Tool zur  
558 quantitativen Landschaftsstrukturanalyse, ESRI Anwenderkonferenz. Innsbruck,  
559 Austria. Available from URL: [http://downloads2.esri.com/campus/uploads/library/  
560 pdfs/68464.pdf](http://downloads2.esri.com/campus/uploads/library/pdfs/68464.pdf) (accessed on 13 September 2012).
- 561 Lindenmayer, D.B, Luck, G., 2005. Synthesis: Thresholds in conservation and  
562 management. *Biol. Conserv.* 124, 351– 354.
- 563 Lopes, L.E., Buzato, S., 2007. Variation in pollinator assemblages in a fragmented  
564 landscape and its effects on reproductive stages of a self-incompatible treelet,  
565 *Psychotria suterella* (Rubiaceae). *Oecologia* 154, 305–314.
- 566 Mabberley, D.J., 1997. The plant book: A portable dictionary of vascular plants.  
567 Cambridge University Press, New York.
- 568 Mahan, C.G.; Yahner, R. H., 1999. Effects of forest fragmentation on behaviour  
569 patterns in the eastern chipmunk (*Tamias striatus*). *Can. J. Zool.*, 77, 1991–97.
- 570 Malcolm, J.R., 1991. The small mammals of Amazonian forest fragments: Pattern and  
571 process. Ph.D. thesis, University of Florida, Gainesville.
- 572 Mcgarigal, K., Marks, B.J., 1995. Fragstats: Spatial pattern analysis program for  
573 quantifying landscape structure. General Technical Report PNW-GTR-351.  
574 Portland, OR: Pacific Northwest Research Station.

- 575 Metzger, J.P., 1999. Estrutura da paisagem e fragmentação: análise bibliográfica. An.  
576 Acad. Bras. Ciênc. 71, 445– 463.
- 577 Metzger, J.P., Décamps, H., 1997. The structural connectivity threshold: an hypothesis  
578 in conservation biology at the landscape scale. Acta Ecol. 18, 1–12.
- 579 Moguel, P., Toledo, V.M., 1999. Biodiversity conservation in traditional coffee systems  
580 of Mexico. Conserv. Biol. 13, 11–21.
- 581 Morellato, L.P.C., 1991. Estudo da fenologia de árvores, arbustos e lianas de uma  
582 floresta semidecídua do sudeste do Brasil. Tese de doutorado, Universidade  
583 Estadual de Campinas, Campinas, São Paulo, Brazil.
- 584 Morreale, S.J., Sullivan, K.L., 2010. Community-level enhancements of biodiversity  
585 and ecosystem services. Front. Earth Sci. China 4, 14–21.
- 586 Murcia, C., 1995. Edge effects in fragment forests: implications for conservation.  
587 Trends Ecol. Evol. 10, 58–62.
- 588 Murcia, C., 1996. Forest fragmentation and the pollination of neotropical plant, in:  
589 Schelhas, J., Greenberg, R. (Eds.), Forest Patches in Tropical Landscapes. Island  
590 Press, Washington, p. 19–36.
- 591 Muriel, S.B., Kattan, G.H., 2009. Effects of Patch Size and Type of Coffee Matrix on  
592 Ithomiine Butterfly Diversity and Dispersal in Cloud-Forest Fragments. Conserv.  
593 Biol. 23, 948–956.
- 594 Niesenbaum, R.A., 1993. Light or pollen– seasonal limitations on female reproductive  
595 success in the understory shrub *Lindera benzoin*. J. Ecol. 81, 315–323.
- 596 Osada, N., Takeda, H., Furukawa, A., Awang, M., 2002. Ontogenetic changes in leaf  
597 phenology of a canopy species, *Elateriospermum tapos* (Euphorbiaceae), in a  
598 Malaysian rain forest. J. Trop. Ecol. 18, 91–105.

- 599 Pietrobon, R.C.V., Paoli, A.A.S., Bieras, A.C., 2011. Morfoanatomia foliar de  
600 *Psychotria hoffmannseggiana* (Willd. ex Roem. & Schult.) Müll. Arg. e *Psychotria*  
601 *trichophora* Müll. Arg. (Rubiaceae). *Naturalia*, 34, 21–42.
- 602 Piñero, D., Sarukhan, J., Alberdi, P., 1982. The costs of reproduction in a tropical palm,  
603 *Astrocaryum mexicanum*. *J. Ecol.* 70, 473–481.
- 604 Pivello, V.R., Metzger, J.P., 2007. Diagnóstico da pesquisa em ecologia de paisagens no  
605 Brasil (2000-2005). *Biota Neotrop.* 7, 21–29.
- 606 Pizo, M.A., Almeida-Neto, M., 2009. Determinants of fruit removal in *Geonoma*  
607 *pauciflora*, an understory palm of neotropical forests. *Ecol. Res.* 24, 1179–1186.
- 608 Prevedello, J.A., Vieira, M.V., 2010. Does the type of matrix matter? A quantitative  
609 review of the evidence. *Biodivers. Conserv.* 19, 1205–1223.
- 610 Rambaldi, D.M., Oliveira, D.A.S., 2005. Fragmentação de ecossistemas: causas, efeitos  
611 sobre a biodiversidade e recomendações de políticas públicas. MMA/SBF, Brasília.
- 612 Ramos, F.N., Santos, F.A.M., 2005. Phenology of *Psychotria tenuinervis* (Rubiaceae) in  
613 Atlantic forest fragments: fragment and habitat scales. *Can. J. Bot.* 83, 1305–1316.
- 614 Ramos, F.N.; Santos, F.A.M., 2006. Floral visitors and pollination of *Psychotria*  
615 *tenuinervis* (Rubiaceae): distance from the anthropogenic and natural edges of an  
616 Atlantic forest fragment. *Biotropica* 38, 383–389.
- 617 Rempel, R., 2010. Patch Analyst for ArcGIS, Centre for Northern Forest Ecosystems  
618 Research, 07.15.2011. Available from URL: [http://flash.lakeheadu.ca/~rrempe/pa](http://flash.lakeheadu.ca/~rrempe/patch/index.html)  
619 [tch/index.html](http://flash.lakeheadu.ca/~rrempe/patch/index.html) (accessed on 13 September 2012).
- 620 Renjifo, L.M., 2001. Effects of natural and anthropogenic landscapes matrices on the  
621 abundance of sub Andean bird species. *Ecol. Appl.* 11, 14–31.

- 622 Ribeiro, C.M., Metzger, J.P., Martensen, A.C., Ponzoni, F.J., Hirota M.M., 2009. The  
623 Brazilian Atlantic Forest: How much is left, and how is the remaining forest  
624 distributed? Implications for conservation. *Biological Conservation*, 142, 1141–  
625 1153.
- 626 Ricketts, T.H., 2004. Tropical Forest Fragments Enhance Pollinator Activity in Nearby  
627 Coffee Crops. *Conserv. Biol.* 18, 1262–1271.
- 628 Rukke, B.A., 2000. Effects of habitat fragmentation: increased isolation and reduced  
629 habitat size reduces the incidence of dead wood fungi beetles in a fragmented forest  
630 landscape. *Ecography* 23, 492–502
- 631 Shapcott, A., 1998. Vagile but inbred: patterns of inbreeding and the genetic structure  
632 within populations of the monsoon rain forest tree *Syzygium nervosum* (Myrtaceae)  
633 in northern Australia. *J. Trop. Ecol.* 14, 595–614.
- 634 Smith-Ramirez, C., Armesto, J.J., 1994. Flowering and fruiting patterns in the temperate  
635 rainforest of Chiloé, Chile: ecologies and climatic constraints. *J. Ecol.* 82, 353–365.
- 636 Soulé, M.E., Alberts, A.C., Bolger, D. T., 1992. The effects of habitat fragmentation on  
637 chaparral plants and vertebrates. *Oikos* 63, 39–47.
- 638 Steffan-Dewenter, I., Münzenberg, U., Bürger, C., Thies, C., Tschardtke, T., 2002.  
639 Scale-dependent effects of landscape context on three pollinator guilds. *Ecology* 83,  
640 1421–1432.
- 641 Swift, T.L., Hannon, S.J., 2010. Critical thresholds associated with habitat loss: a  
642 review of the concepts, evidence, and applications. *Biol. Rev.* 85, 35–53.
- 643 Taylor, C.M., 1996. Overview of the Psychotrieae (Rubiaceae) in the Neotropics. *Opera*  
644 *Bot. Belg.* 7, 261–270.

- 645 Tischendorf, L., Bender, D.J., Fahrig, L., 2003. Evaluation of patch isolation metrics in  
646 mosaic landscapes for specialist vs. generalist dispersers. *Landsc. Ecol.* 18, 41–50
- 647 Turner, M.G., 2005. Landscape ecology: What is the state of the science? *Annu. Rev.*  
648 *Ecol. Evol. Syst.* 36, 319–344.
- 649 Vamosi, J.C., Knight, T.M., Steets, J.A., Mazer, S.J., Burd, M., Ashman, T.L., 2006.  
650 Pollination decays in biodiversity hotspots. *Proc. Natl. Acad. Sci. U S A.* 103, 956–  
651 961.
- 652 Viana, B.F., Boscolo, D., Neto, E.M., Lopes, L.E., Lopes, A.V., Ferreira, P.A., Pigozzo,  
653 C.M., Primo, L.M., 2012. How well do we understand landscape effects on  
654 pollinators and pollination services? *J. Poll. Ecol.* 7, 31– 34.
- 655 Vianello, R.B., Alves, A.R., 1991. *Meteorologia básica e aplicações*. Imprensa  
656 Universitária, Universidade Federal de Viçosa, Viçosa.
- 657 Virgós, E., 2001. Role of isolation and habitat quality in shaping species abundance: a  
658 test with badgers (*Meles meles* L.) in a gradient of forest fragmentation. *J. Biogeogr.*  
659 28, 381– 89
- 660 Virillo, C.B., Ramos, F.N., Castro, C.C., Semir, J., 2007. Floral biology and breeding  
661 system of *Psychotria tenuinervis* Muell. Arg. (Rubiaceae) in the Atlantic rain forest,  
662 SE Brazil. *Acta Bot. Bras.* 21, 879–884.
- 663 Wagenius, S., Lonsdorf, E., Neuhauser, C., 2007. Patch aging and the S-allee effect:  
664 breeding system effects on the demographic response of plants to habitat  
665 fragmentation. *Am. Nat.* 169, 383–397.
- 666 Waser, N.M., Price, M.V., 1991. Outcrossing distance effects in *Delphinium nelsonii*:  
667 pollen loads, pollen tubes, and seed set. *Ecology* 72, 171–179.

- 668 Wesselingh, R.A., 2007. Pollen limitation meets resource allocation: towards a  
669 comprehensive methodology. *New Phytol.* 174, 26–34.
- 670 Wiens, J.A., 2005. Toward a unified landscape ecology. In: Wiens, J., Moss, M. (Eds.),  
671 *Studies in landscape ecology: issues and perspectives in landscape ecology.*  
672 Cambridge University Press, Cambridge. pp. 365-373.
- 673 Wright, S.J., Van Schaik, C.P., 1994. Light and the phenology of tropical trees. *The Am.*  
674 *Nat.* 143, 193–199.
- 675 Zotz, G.; Bermejo, P.; Dietz, H., 1999. The epiphyte vegetation of *Annona glabra* on  
676 Barro Colorado Island Panama. *J. Biogeogr.* 26, 761–776.

677 **Fig. 1.** Conceptual map outlining sampling design and analyses.

678

679 **Fig. 2.** Rank of the sum of percentage of forest and coffee matrix in landscape around  
680 each fragment.

681

682 **Fig. 3.** Conceptual map outlining the relation of landscape structure parameters and  
683 traits of structure of plants and their habitats in the reproduction of forest plants. (+) =  
684 Positive influence. (-) = Negative influence. Pasture: Percentage of pasture matrix.  
685 Coffee: Percentage of coffee matrix. Sugarcane: Percentage of sugarcane matrix. Forest  
686 Area = percentage of native forest. AWMSI= Area Weighted Mean Shape Index  
687 (calculate the irregularity of remaining landscape performing a weighted average for the  
688 area.

689

690 **Table 1:** Landscape and within –patch parameters and location (WGS 1986 23S-UTM)  
691 of fragments studied in the region of Alfenas, MG, Brasil.

692

693 **Table 2:** Ecological hypotheses associated with the GLM used to describe the variation  
694 in flower and fruit production (proportion of annual intensity per individual) of *P.*  
695 *vellosiana* in Atlantic in function of variables (not correlated): CHS (cm), height (plant  
696 height), volume (ellipsoid volume of crown), canopy height (canopy height above the  
697 individual), vertical profile (understory vertical density), DI (distance to the nearest  
698 individual plant) and DC (distance to the nearest plant conspecific individual). The  
699 symbol ~ indicates the relationship of statistical dependence ( $X \sim Y$ , means that the  
700 expected value of Y is a function of X).

701 **Table 3:** List of all models tested.

702

703 **Table 4:** Ecological hypotheses associated with the GLM used to describe the variation  
704 in flower and fruit production (proportion of annual intensity per fragment) of *P.*  
705 *vellosiana* in Atlantic in function of variables (not correlated): PROX (proximity  
706 fragment), Percentage of forest cover, Area Weighted Mean Shape Index and fragment  
707 area (ha). The symbol ~ indicates the relationship of statistical dependence ( $X \sim Y$ ,  
708 means that the expected value of Y is a function of X). The symbol “+” indicates  
709 additive effects. Flower = flower production, fruit = fruit production.

710

711 **Table 5:** Relative importance (GLM) from landscape metrics to total number of floral  
712 visits, number of floral visits of the most frequent species, flower and fruit production,  
713 and the relative importance of features of the plant and its light microenvironment to  
714 flower and fruit production. We presented only the valid models ( $\Delta AICc \leq 2$ ) and the  
715 null models (when the random explains variation).

716

717 **Table 6:** Rank of the number of individuals of each visitor species from *P. vellosiana*  
718 flowers within each fragment (1, 2, 3, 4, 5, 6, 7, 8) in 2011.

719

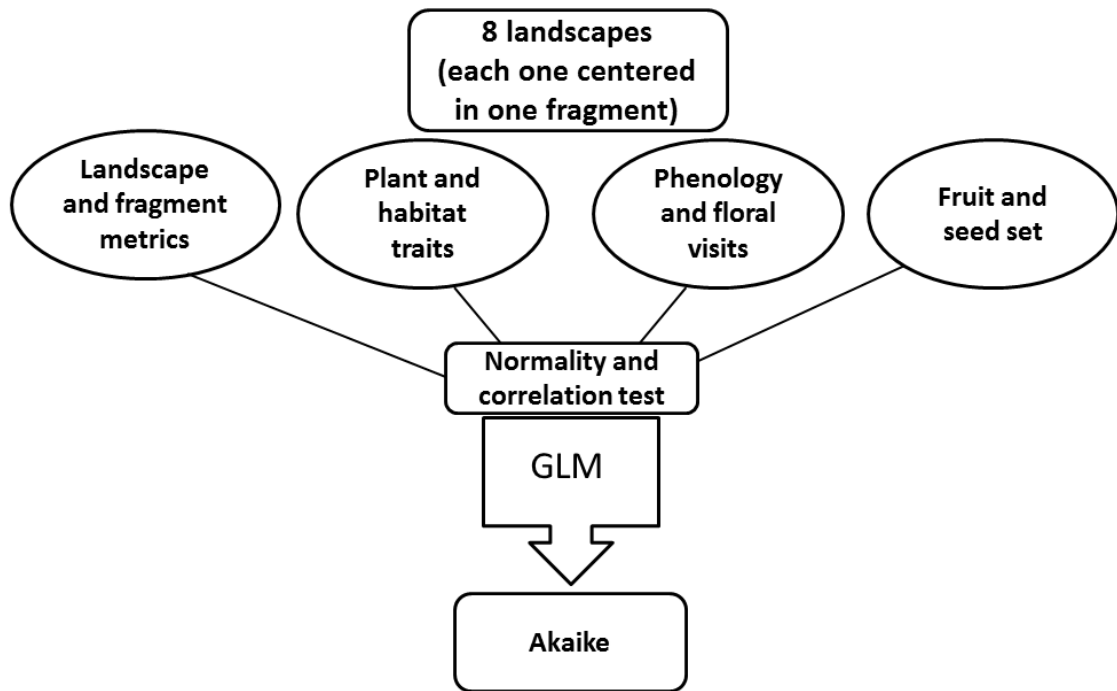
720

721

722

723





724

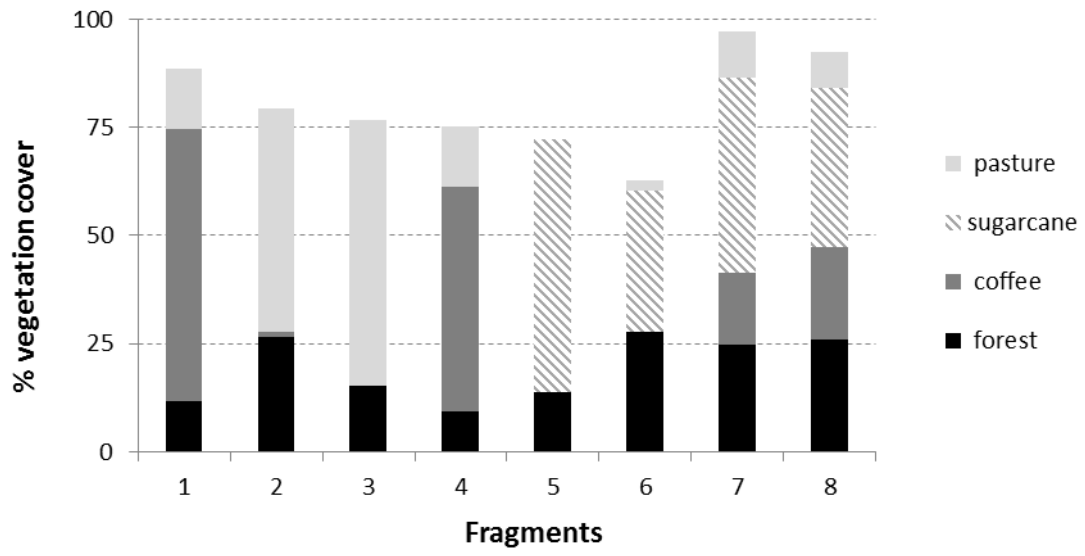
725 **Fig. 1.**

726

727

728

729



730

731

**Fig. 2.**

732

733

734

735

736

737

738

739

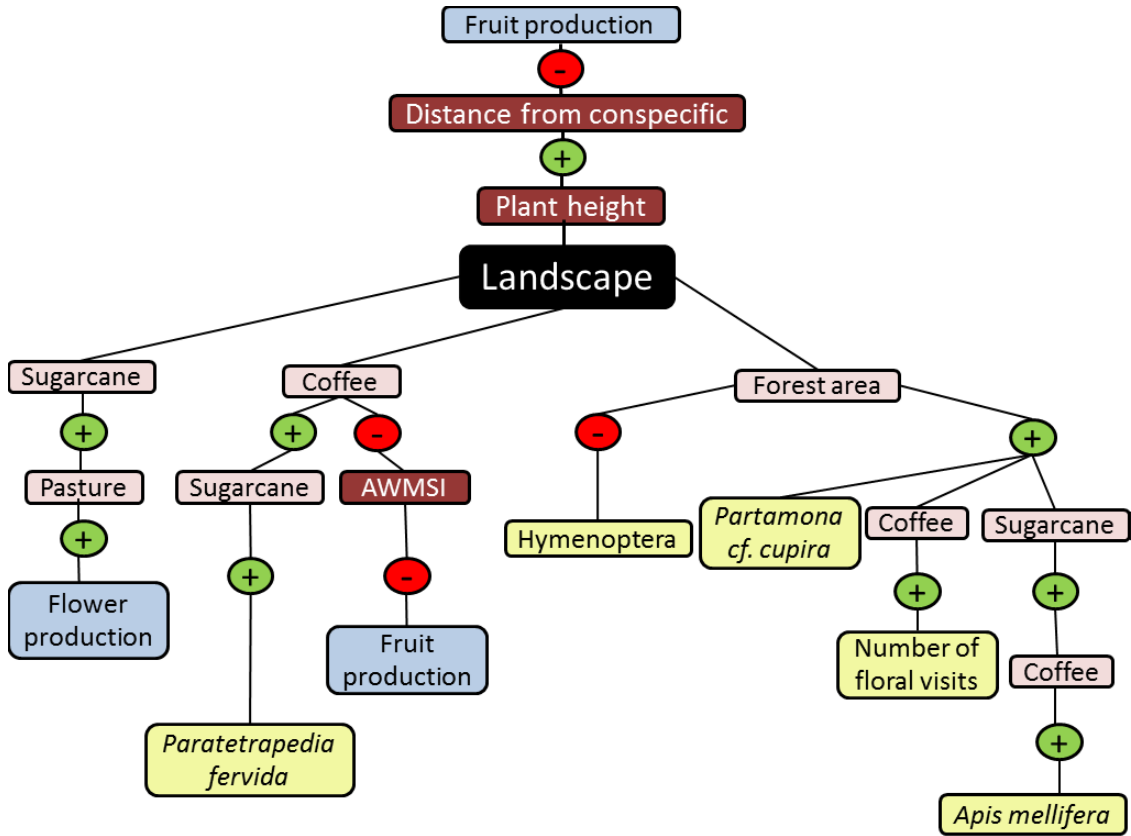
740

741

742

743

744



745

746 **Fig. 3.**

747

748 **Table 1:** Landscape and within –patch parameters and location (SAD 69 23S-UTM) of fragments studied in the region of Alfenas, MG,  
749 Brasil.

<b>Fragment</b>	<b>Regional name</b>	<b>Area (ha)</b>	<b>Longitude</b>	<b>Latitude</b>	<b>% forest</b>	<b>% coffee</b>	<b>% sugarcane</b>	<b>% pasture</b>	<b>AWMSI</b>
1	Paraíso	36.85	411139	7639832	11.73	62.87	0	13.74	1.48
2	Gaspar Lopes	81.55	404019	7635312	26.43	1.17	0	51.56	1.97
3	Matão	46.65	408722	7621882	15.04	0	0	61.51	1.74
4	Cemitério	22.99	403017	7615205	9.38	51.83	0	14.02	1.25
5	I	37.05	386749	386749	13.66	0	58.33	0	1.56
6	Porto	87.18	383793	7630675	27.75	0	32.51	2.3	2.23
7	São José	28.57	381066	7629262	24.54	16.83	45.13	10.5	1.86
8	M	56.05	379160	7626383	25.98	21.19	36.96	8.18	1.93

750  
751 (% forest = percentage of forest cover, % coffee = percentage of coffee matrix, % sugarcane = percentage of sugarcane matrix, % pasture = percentage of pasture matrix, all percentages in function of the  
752 landscape. AWMSI= area weighted mean shape index).  
753

754  
755

756 **Table 2:** Ecological hypotheses associated with the GLM used to describe the variation in flower and fruit production (proportion of  
 757 annual intensity per individual) of *P. vellosiana* in Atlantic in function of variables (not correlated): CSH (cm), height (plant height),  
 758 volume (ellipsoid volume of crown), canopy height (canopy height above the individual), vertical profile (understory vertical density), DI  
 759 (distance to the nearest individual plant) and DC (distance to the nearest plant conspecific individual). The symbol ~ indicates the  
 760 relationship of statistical dependence ( $X \sim Y$ , means that the expected value of Y is a function of X).

Model	Hypothesis	References
1 flower~CSH	Only the CSH interferes with flower production. Higher plant - higher flower production.	Charlesworth and Leon, 1976; Piñero et al., 1982
2 flower~height	Only the plant height interferes with flower production. Higher plant - higher flower production.	Charlesworth and Leon, 1976; Piñero et al., 1982
3 flower~volume	Only the volume of plant crown interferes with flower production. Larger volume of <i>P. vellosiana</i> crown - greater flower production.	Charlesworth and Leon, 1976; Piñero et al., 1982
4 flower~canopy height	Only the canopy height above the <i>P. vellosiana</i> interferes with flower production. Greater canopy height - greater indirect light - greater flower production.	Wright and van Shaik, 1994
5 flower~vertical profile	Only the vertical profile interferes with flower production. Minor vertical profile, greater indirect light, greater flower production.	Wright and van Shaik, 1994

6	flower~DI	Only the distance to the nearest individual plant interferes with flower production. Competition for light and nutrients.	Wright and van Shaik, 1994
7	Flower~DC	Only the distance to the nearest plant conspecific interferes with flower production. Attraction and competition for pollinators.	Duffy and Stout, 2011
8	fruit~CSH	Only the CAS interferes with fruit production. Higher plant - higher fruit production.	Charlesworth and Leon, 1976; Piñero et al., 1982
9	fruit~height	Only the plant height interferes with fruit production. Higher plant - higher fruit production.	Charlesworth and Leon, 1976; Piñero et al., 1982
10	fruit~volume	Only the volume of plant crown interferes with fruit production. Larger volume of <i>P. vellosiana</i> crown - greater fruit production.	Charlesworth and Leon, 1976; Piñero et al., 1982
11	fruit~canopy height	Only the canopy height above the <i>P. vellosiana</i> interferes with fruit production. Indirect effect of luminosity, attraction of pollinators (competition).	Kato and Hiura,1999; Niesenbaum, 1993; Pizo and Almeida-Neto, 2009
12	fruit~vertical profile	Only the vertical profile interferes with fruit production. Indirect effect of luminosity and attraction of pollinators (competition).	Kato and Hiura,1999; Niesenbaum, 1993; Pizo and Almeida-Neto, 2009
13	fruit~DI	Only the distance to the nearest individual plant interferes with fruit production. Indirect effect of luminosity and attraction of pollinators (competition).	Kato and Hiura,1999; Niesenbaum, 1993

- 14 fruit~DC Only the distance to the nearest plant conspecific interferes with fruit production. Indirect effect of luminosity and attraction of pollinators (competition). Niesenbaum, 1993
-

761 **Table 3:** List of all models tested.

	Scale	Dependent Variable	Independent Variable
1	Local (Plant and habitat features)	Flower /Fruit	~CSH
2			~height
3			~volume
4			~canopy height
5			~vertical profile
6			~DI
7			~DC
8			~CSH+height
9			~CSH+vertical profile
10			~CSH+DI
11			~CSH+DC
12			~height+canopy height
13			~height+vertical profile
14			~height+DI
15			~height+DC
16			~volume+canopy height
17			~volume+vertical profile
18			~volume+DI
19			~volume+DC
20			~canopy height+ vertical profile
21			~canopy height+DI
22			~canopy height+DC



23			~vertical profile+DI
24			~vertical profile +DC
25			~DI+DC
<hr/>			
26	Landscape	Flower	~% forest
27			~% coffee
28			~% sugarcane
29			~% pasture
30			~% forest +% coffee
31			~% forest +% sugarcane
32			~% forest + % pasture
33			~ % Café + % cana
34			~% coffee+% pasture
35			~% sugarcane +% pasture
36			~% AWMSI
37			~% AWMSI+% coffee
38			~% AWMSI+% sugarcane
39			~% AWMSI+% pasture
<hr/>			
40	Landscape	Fruit, fruit set, seed set,	~prox
41		number of	~% forest
42		total visits,	~% coffee
43		richness of	~% sugarcane
44		total visits,	~% pasture
45		number of	~prox+% forest
46		visits of	~prox+% coffee
		species A, B,	
		C, D and E.	

47	~prox+sugarcane
48	~prox+pasture
49	~% forest +% coffee
50	~% forest +% sugarcane
51	~% forest + % pasture
52	~ % coffee + % cane
53	~% coffee+% pasture
54	~% sugarcane +% pasture
55	~% AWMSI
56	~% AWMSI+prox
57	~% AWMSI+% coffee
58	~% AWMSI+% sugarcane
59	~% AWMSI+% pasture

762  
763  
764  
765  
766  
767  
768

---

(species of floral visitors: A= *Partamona cf. cupira*, B= Hymenoptera, C= *Paratetrapedia fervida*, D= Syrphidae, E= *Apis mellifera*. CSH= height stem circumference, height (plant height), volume (ellipsoid volume of crown), canopy height (canopy height above the individual), vertical profile (understory vertical density), DI (distance to the nearest individual plant), DC (distance to the nearest plant conspecific individual), % forest = percentage of forest cover, % coffee = percentage of coffee matrix, % sugarcane = percentage of sugarcane matrix, % pasture = percentage of pasture matrix, all percentages in function of the landscape. AWMSI = area weighted mean shape index).

769 **Table 4:** Ecological hypotheses associated with the GLM used to describe the variation in flower and fruit production (proportion of  
 770 annual intensity per fragment) of *P. vellosiana* in Atlantic in function of variables (not correlated): PROX (proximity fragment), Percentage  
 771 of forest cover, Area Weighted Mean Shape Index and fragment area (ha). The symbol ~ indicates the relationship of statistical dependence  
 772 ( $X \sim Y$ , means that the expected value of Y is a function of X). The symbol “+” indicates additive effects. Flower = flower production, fruit  
 773 = fruit production.

	Model	Hypothesis	References
1	flower~ %forest	Only the percentage of forest cover interferes with the flower production. Increase of forest cover increases the flower production.	Lindenmayer and Luck, 2005
2	flower~%coffee	Only the percentage of coffee matrix interferes with the flower production. Increase of coffee matrix increases the flower production.	Prevedello and Vieira, 2010
3	flower~%sugarcane	Only the percentage of sugarcane matrix interferes with the flower production.	Prevedello and Vieira, 2010
4	flower~%pasture	Only the percentage of pasture matrix interferes with the flower production.	Prevedello and Vieira, 2010
5	flower ~ AWMSI	Only the area weighted mean shape index interferes with the flower production.	McGarigal and Marks, 1995; Murcia, 1995

6	fruit~PROX	Only the proximity fragment interferes with fruit production. Increase of proximity fragment increases fruit production.	Farwig et al., 2009; Ferreira et al., 2013
7	fruit~%forest	Only the percentage of forest cover interferes with fruit production. Increase of forest cover increases fruit production	Lindenmayer and Luck, 2005
8	fruit~%coffee	Only the percentage of coffee matrix interferes with fruit production. Increase of coffee matrix increases fruit production.	Prevedello and Vieira, 2010
9	fruit~%sugarcane	Only the percentage of sugarcane matrix interferes with fruit production.	Prevedello and Vieira, 2010
10	fruit~%pasture	Only the percentage of pasture matrix interferes with fruit production.	Prevedello and Vieira, 2010
11	fruit~AWMSI	Only the area weighted mean shape index interferes with fruit production. Increase of area weighted mean shape index decreases fruit production.	McGarigal and Marks, 1995; Murcia, 1995
12	N_visits~PROX	Only the proximity fragment individual interferes with the number of floral visits. Increase of proximity fragment increases fruit production	Farwig et al., 2009; Ferreira et al., 2013
13	N_visits ~%forest	Only the percentage of forest cover interferes with the number of floral visits. Increase of forest cover increases the number of floral visits.	Lindenmayer and Luck, 2005

14	N_visits ~%coffee	Only the percentage of coffee matrix interferes with the number of floral visits. Increase of coffee matrix increases the number of floral visits.	Prevedello and Vieira, 2010
15	N_visits ~%sugarcane	Only the percentage of sugarcane matrix interferes with the number of floral visits.	Prevedello and Vieira, 2010
16	N_visits ~%pasture	Only the percentage of pasture matrix interferes with the number of floral visits.	Prevedello and Vieira, 2010
17	N_visits ~AWMSI	Only the area weighted mean shape index interferes with the number of floral visits	McGarigal and Marks, 1995; Murcia 1995

774  
775  
776

---

(% forest = percentage of forest cover, % coffee = percentage of coffee matrix, % cane= percentage of sugarcane matrix, % pasture = percentage of pasture matrix, all percentages in function of the landscape.  
AWMSI= area weighted mean shape index, plant height= total height of the individual of *P. vellosiana* and distance conspecific= distance from the nearest conspecific).

777 **Table 5:** Relative importance (GLM) from landscape metrics to total number of floral visits, number of floral visits of the most frequent  
 778 species, flower and fruit production, and the relative importance of features of the plant and its light microenvironment to flower and fruit  
 779 production. We presented only the valid models ( $\Delta AICc \leq 2$ ) and the null models (when the random explains variation).

Dependent variable	Independent variable	Equation	AICc	$\Delta AICc$	wAICc
Total number of floral visits	% forest + % coffee	$Y = 1.1 + 0.0003x + 0.0006x_1$	76.84	0.00	0.46
	% coffee	$Y = 2.4 + 0.0001x$	77.23	0.39	0.38
	None		89.82	12.98	0.00
Flower	% sugarcane + % pasture	$Y = 5.8 + 0.0001x + 0.00005x_1$	223.32	0.00	1.00
	None		406.57	183.25	0.00
Ripe fruit	AWMSI + % coffee	$Y = 10.98 - 0.03x - 0.0002x_1$	714.31	0.00	1.00
	None		1714.40	1000.09	0.00
Ripe fruit	Plant height x distance conspecific	$Y = 0.5 + 0.004x - 0.0008x_1$	1093.33	0.00	0.99
	None		1170.80	77.47	0.00
<i>Partamona cf. cupira</i>	% forest	$Y = 0.7 + 0.0003x$	44.68	0.00	0.95
	None		89.82	45.14	0.00

Hymenoptera	% forest	$Y = 6.4 - 0.005x$	44.68	0.00	0.95
	None		89.82	45.14	0.00
<i>Paratetrapedia fervida</i>	% coffee + % sugarcane	$Y = -218 + 0.044x + 0.045x_1$	26.94	0.00	0.96
	None		60.55	33.61	0.00
<i>Apis mellifera</i>	% forest	$Y = -1.39 + 0.0004x$	26.22	0.00	0.67
	% sugarcane + % pasture	$Y = -63.6 + 0.01x + 0.01x_1$	27.94	1.72	0.28
	None		38.37	12.15	0.00

780

781 (% forest = percentage of forest cover, % coffee = percentage of coffee matrix, % sugarcane = percentage of sugarcane matrix, % pasture = percentage of pasture matrix, all percentages in function of the

782 landscape. AWMSI= area weighted mean shape index, plant height = total height of the individual of *P. vellosiana* and distance conspecific = distance from the nearest conspecific).

783

784

785

786

787

788 **Table 6:** Rank of the number of individuals of each visitor species from *P. vellosiana* flowers within each fragment (1, 2, 3, 4, 5, 6, 7, 8) in  
 789 2011.  
 790

Order	Species	Fragments								Total per specie
		1	2	3	4	5	6	7	8	
Hymenoptera	<i>Partamona cf. cupira</i> (Apidae)	0	6	2	5	1	2	5	22	43
Hymenoptera	Hymenoptera sp	21	0	1	1	2	1	0	0	26
Hymenoptera	<i>Paratetrapedia fervida</i> (Anthophoridae)	12	0	0	0	0	0	4	0	16
Hymenoptera	<i>Apis mellifera</i> (Apidae)	0	2	0	0	1	0	0	7	10
Diptera	Syrphidae sp	1	0	0	1	4	2	0	2	10
Hymenoptera	<i>Augochlora sp</i> ( <b>Halictidae</b> )	0	1	0	1	2	1	0	2	7
Hymenoptera	<i>Melipona quadrifasciata</i> (Apidae)	0	0	2	2	1	2	0	0	7
Lepidoptera	Lepidoptera sp1	0	3	0	0	0	0	0	0	3
Lepidoptera	Lepidoptera sp2	0	0	0	3	0	0	0	0	3
Diptera	Diptera sp	0	1	0	0	0	0	0	1	2



Hymenoptera	<i>Trigona spinipes</i> (Apidae)	0	0	0	0	0	0	1	0	1
Lepidoptera	Lepidoptera sp3	0	0	0	0	1	0	0	0	1
Lepidoptera	Lepidoptera sp4	0	0	0	0	1	0	0	0	1
Lepidoptera	Lepidoptera sp5	0	1	0	0	0	0	0	0	1
Total per fragment		34	14	5	13	13	8	10	34	131

791

792

---

(Total visits= sum of the number of individuals of each visitor species of all fragments, Total per fragment = sum of the number of individuals of all visitor species per fragment).

#### **4 CONSIDERAÇÕES FINAIS**

Podemos concluir com este trabalho que variáveis da estrutura da paisagem (quantidade de habitat ou % de cobertura de floresta nativa, tipo de matriz e forma do fragmento) e de características locais, relacionadas à planta e seu habitat (grau de agregação da população de plantas e tamanho dos seus indivíduos) influenciam a reprodução de plantas florestais em locais onde a perda de habitat já alcançou um nível crítico. Essas informações são importantes porque geram conhecimento teórico e prático que podem ser utilizados em orientações ao produtor rural, tomadores de decisão e políticas públicas quanto ao uso da terra, conservação e restauração das áreas degradadas. Um exemplo, é o guia destinado a produtores rurais (Figura 1), criado por nós, como a finalidade de divulgar à sociedade algumas informações geradas e abordadas por este trabalho.

Alguns termos importantes da Ecologia de Paisagem foram descritos (Anexo A) como a finalidade de facilitar a leitura por pessoas de outras áreas, além disso, diagramas das relações que apresentaram resultados relevantes (Anexo B) (melhores modelos para explicar os resultados) são apresentados como material suplementar da dissertação.

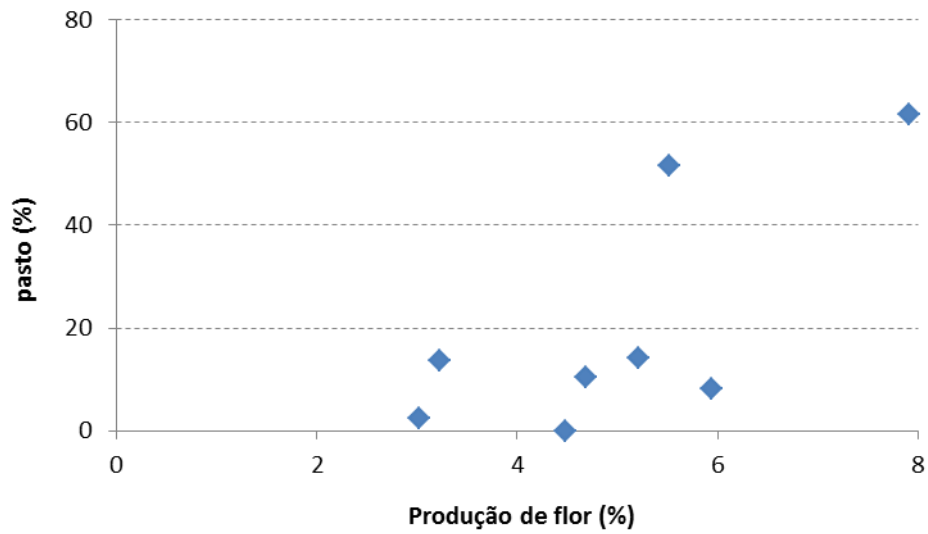
**Figura 1.** Guia de educação ambiental que pode ser destinado ao produtor rural, como forma de repassar conhecimentos abordados e gerados por esse trabalho.



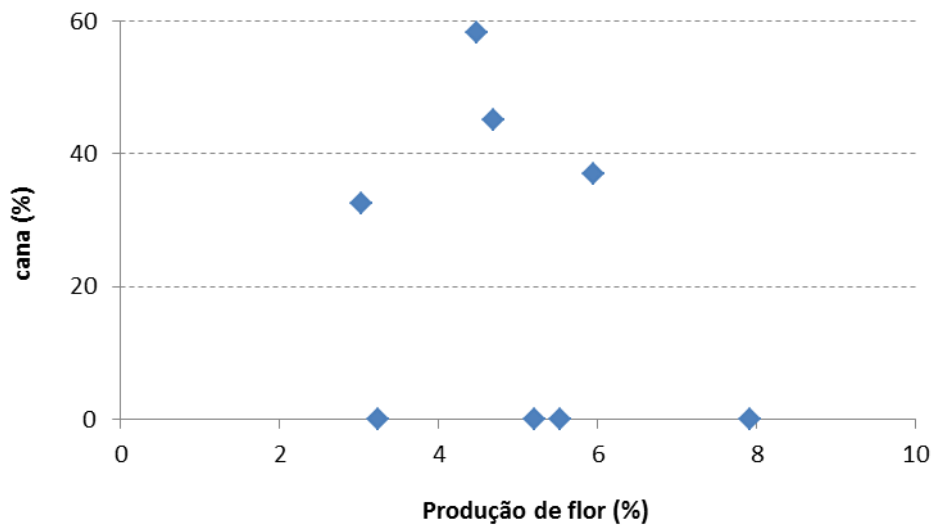
## ANEXO A - Termos importantes da Ecologia de Paisagem

A ecologia de paisagem alterou e incorporou termos ecológicos já existentes e criou novos termos, muitos desses tão interconectados como a própria área. O entendimento desses termos tem fundamental importância para a compreensão dessa nova abordagem de ecologia e entre os mais usados pode-se citar: **escala** (dimensão espacial ou temporal de processos ou objetos, ou quantidade de resolução espacial), **paisagem** (áreas geográficas espacialmente heterogêneas, caracterizadas por diversas interações de ecossistemas, desde sistemas aquáticos e terrestres relativamente naturais como as florestas, campos e lagos, até ambientes dominados pelo homem, incluindo cenários urbanos e agrícolas), **conectividade** (continuidade espacial de um habitat ou tipo de cobertura que atravessa uma paisagem), **heterogeneidade** (medida do quanto as partes de uma paisagem se diferenciam de outra ou distribuição aleatória de objetos através da paisagem), **mancha** (área homogênea, restrita e não linear da paisagem que se distingue das unidades vizinhas), **corredores** (estruturas lineares da paisagem que diferem das unidades vizinhas e que ligam pelo menos dois fragmentos de habitat anteriormente unidos), **matriz** (unidade da paisagem funcionalmente e espacialmente dominante), **bordas** (áreas de transição ou de contato brusco entre diferentes unidades da paisagem), **fragmentação** (quebra de um habitat, ecossistema ou tipo de uso do solo em parcelas menores) (METZGER, 2001; TURNER; GARDNER; O'NEILL, 2001; WU, 2007).

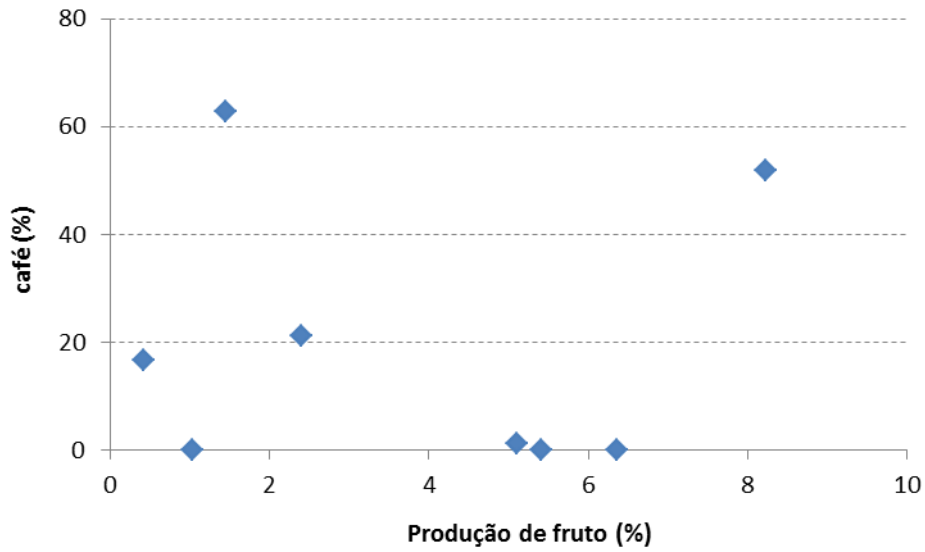
**ANEXO B** – Diagramas de dispersão da influência de parâmetros estruturais de paisagem e de escala local sobre respostas reprodutivas das plantas.



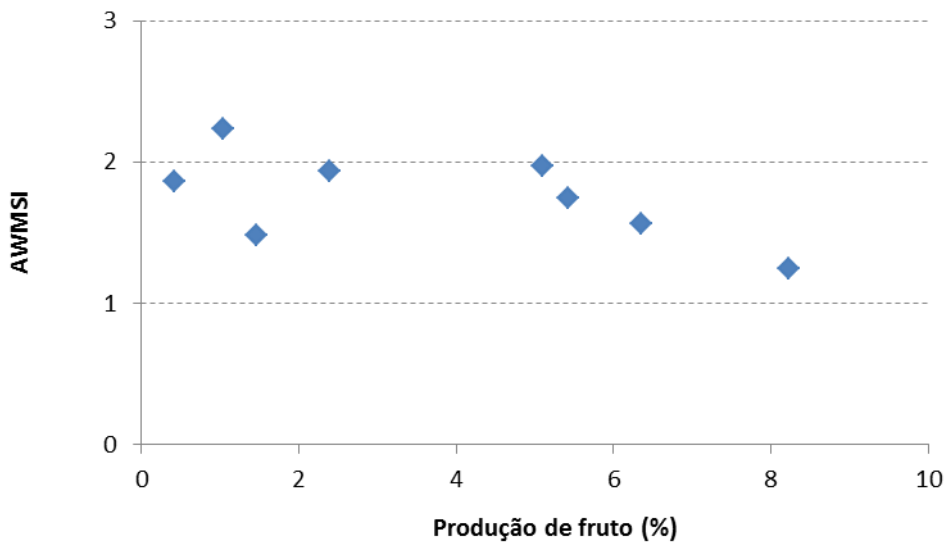
**Figura 1** – Diagrama de dispersão da relação entre cobertura da paisagem por matriz de pasto (%) e produção de flor (%).



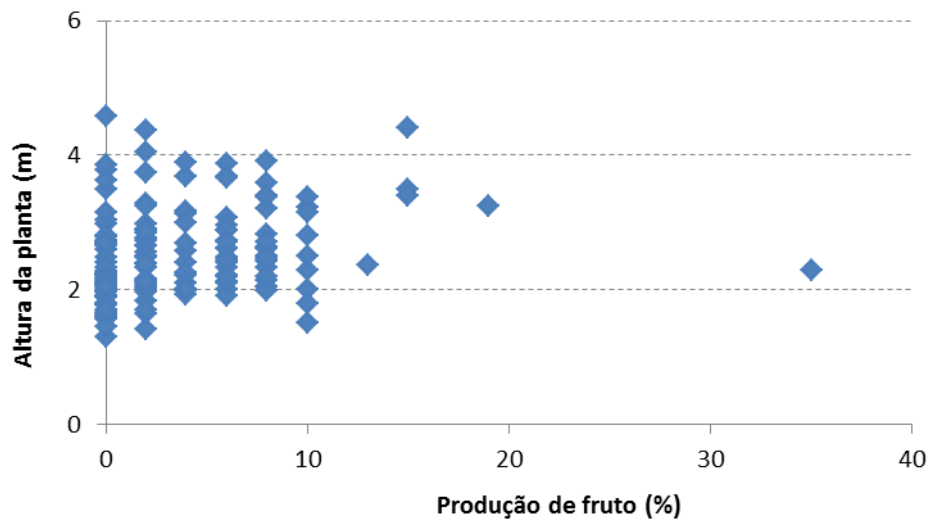
**Figura 2** – Diagrama de dispersão da relação entre cobertura da paisagem por matriz de cana (%) e produção de flor (%).



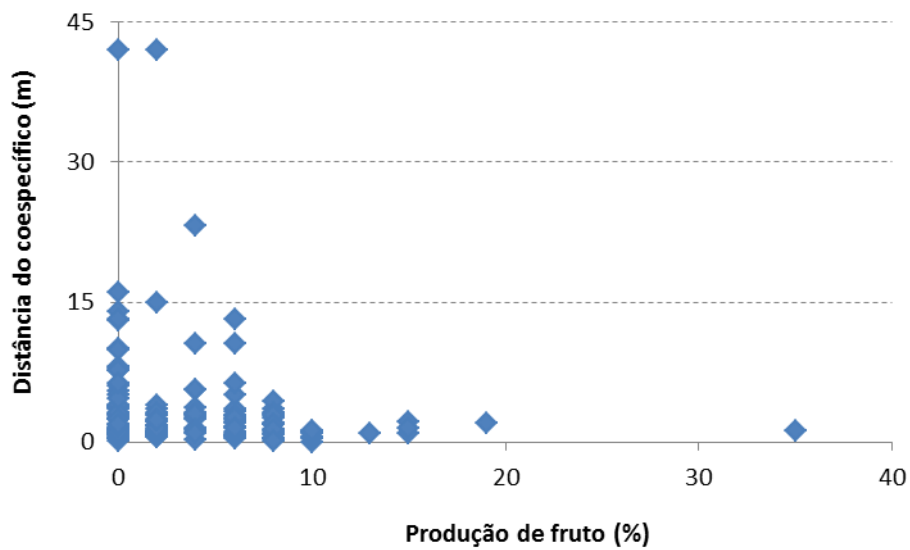
**Figura 3** – Diagrama de dispersão da relação entre cobertura da paisagem por matriz de café (%) e produção de fruto (%).



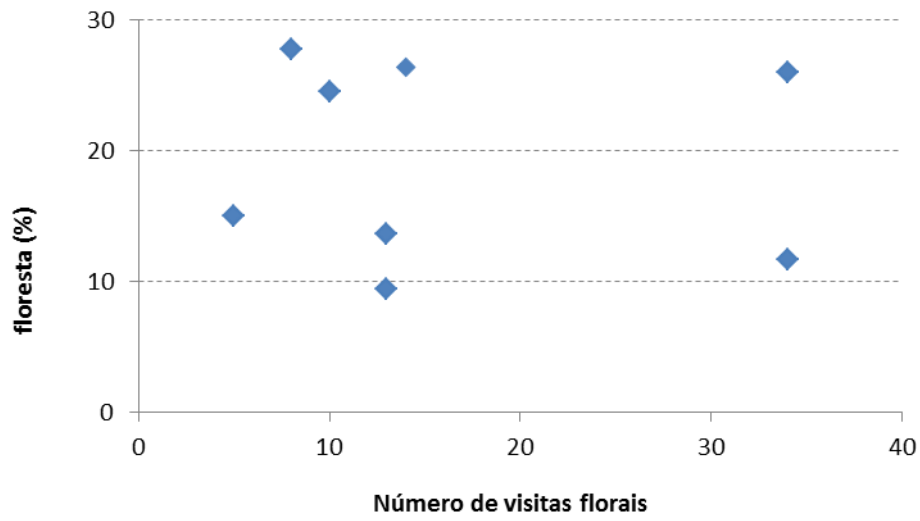
**Figura 4** – Diagrama de dispersão da relação da forma ponderada pela área das manchas na paisagem (AWMSI) e produção de fruto (%).



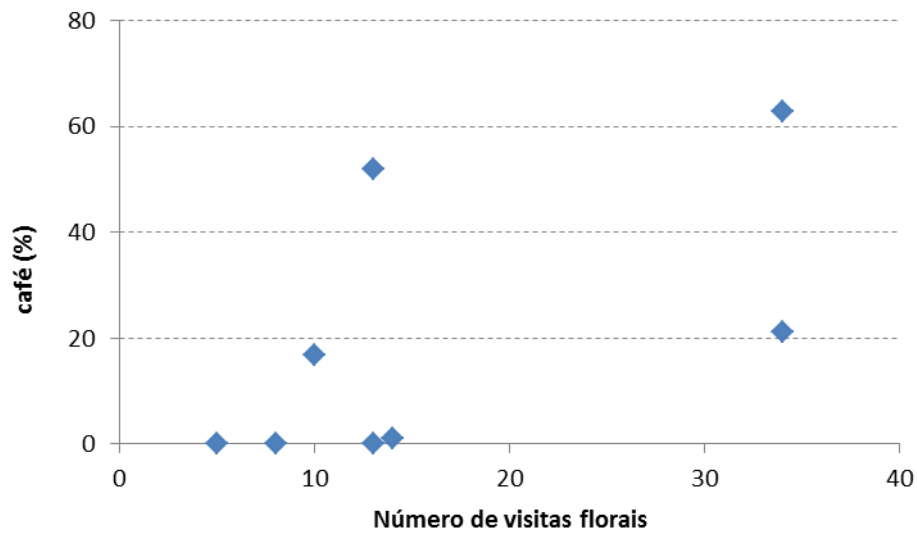
**Figura 5** – Diagrama de dispersão da relação entre a altura dos indivíduos de *Psychotria vellosiana* (m) e produção de fruto (%).



**Figura 6** – Diagrama de dispersão da relação entre distância do coespecífico mais próximo da *Psychotria vellosiana* e produção de fruto (%).



**Figura 7**– Diagrama de dispersão da relação entre cobertura da paisagem por floresta nativa (%) e número de visitas florais.



**Figura 8**– Diagrama de dispersão da relação entre cobertura da paisagem por matriz de café (%) e número de visitas florais.